

# Recommended Procedures for Toxic Air Contaminant Health Risk Assessments

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## **Cleaner Air Oregon Program**

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DEQ is a leader in restoring, maintaining and enhancing the quality of Oregon's air, land and water.



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## LIST OF ABBREVIATIONS

ABC	DEQ Ambient Benchmark Concentration
ADAF	Age Dependent Adjustment Factor
AERMOD	American Meteorological Society/EPA preferred air dispersion modeling program
AERSCREEN	Program to run AERMOD in screening mode
ATSAC	DEQ's Air Toxics Science Advisory Committee
ATSDR	Agency for Toxic Substances and Disease Registry
CAO	Cleaner Air Oregon
CDDs/CDFs	Chlorinated Dibenzo- <i>p</i> -dioxins and Chlorinated Dibenzofurans
CSM	Conceptual Site Model
DEQ	Oregon Department of Environmental Quality
DF	Dispersion Factor
ED	Exposure Duration
ELAF	Early-Life Adjustment Factor
EPA	U.S. Environmental Protection Agency
ER	Emission Rate
ESRE	Emission Scaled Risk Estimate
EQC	Environmental Quality Commission
HI	Hazard Index
HQ	Hazard Quotient
IRIS	EPA's Integrated Risk Information System
IUR	Inhalation Unit Risk
LRAPA	Lane Regional Air Protection Agency
MAKEMET	Program that generates a site-specific matrix of meteorological conditions for input to AERMOD or AERSCREEN
MPAF	Multipathway Adjustment Factor
MRL	ATSDR's Minimal Risk Level
NRAF	Nonresident Adjustment Factor
NSR	New Source Review
OAR	Oregon Administrative Rules
OEHHA	California's Office of Environmental Health Hazard Assessment
OHA	Oregon Health Authority
OSHA	Occupational Safety and Health Administration
PAH	Polycyclic Aromatic Hydrocarbon
PBT	Persistent, Bioaccumulative, and Toxic air contaminant
PCB	Polychlorinated Biphenyl
PPRTV	EPA's Provisional Peer-Reviewed Toxicity Value
RAL	Risk Action Level
RBC	Risk-Based Concentration
RBDM	Risk-Based Decision Making
RDR	Risk Determination Ratio
REER	Risk Equivalent Emission Rate
RfC	Reference Concentration
SCAQMD	California's South Coast Air Quality Management District
TBACT	Toxics Best Available Control Technology
TAC	Toxic Air Contaminant
TEF	Toxic Equivalency Factor
TEU	Toxics Emission Unit
TRV	Toxicity Reference Value

# 1. INTRODUCTION

## 1.1 Purpose and Organization

This document provides DEQ's recommended procedures for conducting toxic air contaminant (TAC) risk assessments in compliance with OAR chapter 340, division 245. A risk assessment can range from a simple risk assessment using screening-type modeling (Levels 1 and 2), to a risk assessment requiring approved air dispersion models (Level 3), or finally a complex risk assessment (Level 4) that may include site-specific adjustments to exposure assumptions.

The methods to perform human health risk assessments at sources that emit toxic air contaminants in Oregon are based primarily on U.S. Environmental Protection Agency guidance (EPA 1989), and are consistent with and make reference to human health risk assessment guidance under DEQ's Cleanup Program (DEQ 2010). In general, the exposure factors and equations described in this document are sufficient for calculating exposure and risk from existing, modified, reconstructed, and new facilities.

Section 2 outlines general risk assessment concepts, including the development of the Risk Based Concentrations (RBCs) DEQ will use to assess risk from a facility. This section also provides an overview of the risk assessment process, including development of a conceptual site model, and a brief discussion on air dispersion modeling. Section 3 presents detailed discussion on conducting a risk assessment, and provides calculation methodologies and suggested guidance for each level of risk assessment. The appendices contain walk-through examples of risk assessment calculations at different levels and using different approaches, as well as a discussion of the development of Multipathway Adjustment Factors and RBCs. Also contained in the appendices are tables of target organs for use in noncancer risk evaluations.

DEQ intends this document to serve as a guide for facilities preparing risk assessments for the Cleaner Air Oregon program by outlining approvable, recommended procedures. Additionally, DEQ strongly encourages facilities to establish and maintain an open and collaborative communication with DEQ during this process to ensure an efficient and successful risk assessment.

## 1.2 Process Overview

The overall recommended human health risk assessment process involves the general steps discussed below. Information on existing site conditions and the nature of properties potentially impacted by facility emissions are key prerequisites for screening steps and risk assessments.

All levels of risk assessment involve calculating three separate risk numbers: cancer, chronic noncancer, and acute noncancer. These calculations use the Risk-Based Concentrations (RBCs) listed in OAR 340-245-8010 Table 2. Risks associated with individual toxic air contaminants are summed and then compared with the Risk Action Levels (RALs) in OAR 340-245-8010 Table 1 to determine what action is needed. Different RALs apply to new/reconstructed and existing facilities.

The owner or operator of the source can choose to start with any level of risk assessment, and is not required to do all four levels. While each of the four levels is considered a risk assessment, Levels 1 and 2 function more as risk screening assessments that can be further refined as desired by a Level 3

or 4 risk assessment. More information about these risk assessment levels is included in Section 3. The elements of the different levels of evaluation are the following:

**Level 1.** This risk assessment level involves choosing dispersion factors from OAR 340-245-8010 Table 3 based on site-specific information. This includes stack height and distances to various exposure locations for stack emissions, and building height, dimensions, and distances to exposure locations for fugitive emissions. In the absence of site-specific information, you can use a default dispersion factor. The default dispersion factor is in the upper left-hand corner of each Table 3 A, B, C, and D, and uses the most conservative assumptions (lowest emission release height and closest exposure location). To screen your emissions, multiply each toxic air contaminant emission rate by the dispersion factor, and compare the resulting calculated ambient air concentration with the appropriate RBC in OAR 340-245-8010 Table 2 for residential and non-residential exposure locations and acute exposure locations. Finally, compare the summed excess cancer risks and hazard indices with the RALs in OAR 340-245-8010 Table 1.

**Level 2.** At this level, you can use site-specific information (such as stack height, other stack parameters, and distances to various exposure locations) and perform simple modeling using EPA's AERSCREEN model or AERMOD model in screening mode to calculate ambient air concentrations for comparison with RBCs.

**Level 3.** At this level, you use detailed site-specific information (such as stack heights, building heights, topography, and distances to various exposure locations) and site-specific meteorological data to perform complex modeling using EPA's AERMOD model to calculate ambient air concentrations for comparison with RBCs.

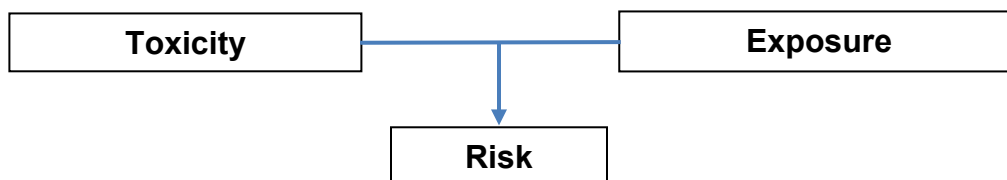
**Level 4.** The most comprehensive risk assessment option uses the same air dispersion modeling conducted in Level 3, with detailed site-specific information. In addition, you can consider factors to refine the exposure assessment. These factors can include modified exposure assumptions, relative bioavailability of toxic air contaminants, or multipathway considerations not covered by the values used to develop RBCs.

## 2. RISK ASSESSMENT OVERVIEW

### 2.1 Risk Assessment Concepts

The goal of the Cleaner Air Oregon program is to evaluate risk to people near facilities that emit regulated toxic air contaminants in OAR 340-247-8010 Table 1, and manage risk consistent with established Risk Action Levels and related regulatory requirements

Risk considers both exposure and toxicity:



Exposure is how much contact someone has with a toxic air contaminant. This mainly includes the concentration of the toxic air contaminant in air, typically expressed as micrograms of toxic air contaminant per cubic meter of air ( $\mu\text{g}/\text{m}^3$ ). The greater the concentration of a toxic air contaminant in air, the greater the risk. Other considerations for exposure include how long the exposure occurs, which for chronic exposure includes both exposure frequency, such as 8 hours per day for workers, and exposure duration, such as 25 years. Acute effects are evaluated using 24-hour averages for exposure to a toxic air contaminant at locations where people may spend several hours of one day (OAR 340-245-0020(3) and (4)).

Toxicity is a measure of how harmful a toxic air contaminant is if someone is exposed to it. The two general types of toxic effects, noncancer and cancer, are evaluated separately. For noncancer effects, we assume there is a threshold below which toxic effects are unlikely to occur. This level is called a reference concentration (RfC).

For cancer effects, the assumption is that there is no threshold for adverse effects. Although the risk at a very low concentration of a carcinogen may be very low, we assume it is not zero. Because of this assumption, the toxicity of carcinogens is given not as a threshold concentration, but instead as a probability of getting cancer as a result of being exposed continuously to a concentration of  $1 \mu\text{g}/\text{m}^3$ . This value is called the Inhalation Unit Risk (IUR) value. For ease of use in assessing risk in the Cleaner Air Oregon program, DEQ converted IURs to concentrations using a target excess cancer risk level of one in one million.

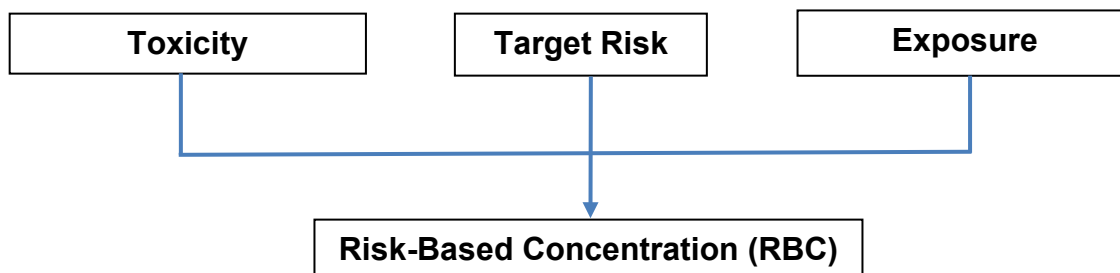
Roughly speaking, if one million people are exposed to an excess cancer risk of one in one million, we would expect about one additional cancer in the population, compared with the already-existing nationwide background level of approximately 400,000 cancers per million people (NIH 2020). We expect the number of people exposed to toxic air contaminants from a single facility to be far less than one million, so the calculated excess cancers in the exposed population (called a cancer burden) as a result of emissions from a facility is expected to be much less than 1. To be clear, in the Cleaner Air Oregon program, DEQ looks at individual probabilities resulting from exposure to toxic air contaminants at specific exposure locations, and not a total population cancer burden.

Toxic air contaminants may have both noncancer and cancer effects. As a general term, we use Toxicity Reference Value (TRV) to mean either the noncancer reference concentration, or a



concentration based on the cancer inhalation unit risk value. Toxicity reference values only consider risks from direct inhalation of toxic air contaminants in air. They do not consider risks from cross-media exposure, such as eating vegetables grown in soil where toxic air contaminants settled out of air into the soil and were taken up into the vegetables. To address cross-media risk in the development of RBCs, an adjustment is made to certain toxicity reference values as described in Section 2.5, and presented in Table C-1 in Appendix C. Other adjustments to TRVs are also discussed in Section 2.5.

DEQ developed RBCs for each toxic air contaminant using standard exposure and toxicity assumptions, generally from EPA, and selected risk levels discussed below. DEQ developed RBCs using the same toxicity and exposure information used in a risk assessment, along with target risk levels. There are separate RBCs for cancer risk, chronic noncancer risk, and acute noncancer risk.



For establishing RBCs for noncarcinogens, target risk is set at a hazard quotient, or HQ, of 1. A hazard quotient is the ratio of the concentration of toxic air contaminant in air to the RBC. An HQ below 1 means there is little likelihood that even sensitive people will experience serious adverse health effects. To establish RBCs for carcinogens, target risk is set at an excess cancer risk of one in one million.

The choices of a HQ of one and an excess lifetime cancer risk level of one in one million are for convenience in establishing RBCs. These risk levels are not necessarily the same as acceptable risk levels, or RALs. OAR 340-245-8010 Table 1 shows RALs, which are the levels at which facilities must take action to address risk. DEQ developed separate RALs for new/reconstructed and existing facilities.

## 2.2 Risk Assessment Process

A number of elements are important to conducting a risk assessment. The first risk assessment elements are important even at a simple evaluation level. Some of the later elements are important only in a Level 4 risk assessment.

The final risk assessment summary report should include the following:

- A problem formulation step to determine a conceptual site model (CSM) describing toxic air contaminant releases and relevant exposure scenarios based on current exposure populations;
- A toxicity analysis evaluating the inherent toxicity of toxic air contaminants;
- An exposure analysis, which includes quantifying exposure concentrations based on the CSM, selecting exposure models, and selecting exposure factor values;
- A risk characterization combining the results of the toxicity and exposure analyses to evaluate risk; and
- A quantitative or qualitative uncertainty analysis covering all aspects of the risk assessment.

The risk assessment approach presented in CAO rules uses a comparison with RBCs, rather than the traditional method where risk is calculated without benefit of RBCs (EPA 1989). A Level 4 risk assessment may use a combined approach for pathways not covered by RBCs in rule.

If you document the risk assessment results in a clear and consistent manner, it will be easier for DEQ staff to review it quickly. To further expedite review of the risk assessment, we recommend that you provide DEQ with electronic copies of spreadsheets of data and calculations with functioning (unlocked) formulas as part of the documentation. If any information is confidential business information (CBI), it can be labeled as such and will be treated following all required procedures to protect the information. DEQ may require some, or all, of this information if the results of the risk assessment cannot be verified with the information provided.

## 2.3 Conceptual Site Model

It is important to have a detailed understanding of the locations and configurations of all toxic air contaminant Toxics Emissions Units (TEUs), exposure pathways, routes of exposure, and types of exposure locations near your facility. A good way of presenting a conceptual site model (CSM) is in a chart, although for most air emission evaluations a brief narrative should be sufficient. If your facility emits persistent, bioaccumulative, and toxic (PBT) air contaminants, and therefore may be required to conduct a multipathway risk assessment, we recommend you describe the site with a more extensive CSM. Figure 3 provides an example of a multipathway conceptual site model.

The conceptual site model should include a list of chemicals being emitted from the facility. This list is derived from the emissions inventory, which is the first technical submittal required under the CAO program. DEQ provides assistance on how an emission inventory should be prepared and presented.

A high-quality CSM should combine information on toxic air contaminants, exposure locations, and exposure pathways to summarize relevant site information for use in the risk assessment. If land is zoned for uses allowing residents, include residential exposure, except as described below. You should consider likely exposure scenarios based on the intended activity at a location. For example, in farmland where a residence is allowed, include exposure to any current houses. However, it is not necessary to consider an unlikely future addition of a house in an agricultural field. Zoning information can be the starting point for this analysis. Zoning is available at the state, local, or city level. DEQ uses state zoning. If more local data is available, provide the appropriate reference so DEQ can verify the information.

For areas zoned only for residential use, evaluate residential exposure to the entire area. If you know that people do not actually live in the area, you can provide documentation to DEQ that there is no current residential use in the area. If DEQ concludes that the documentation is adequate to rebut the presumption of residential use, you can adjust your exposure assessment accordingly. However, you must annually demonstrate that the excluded zoned areas continue to not be used in the manner allowed by the land use zoning.

Knowing how nearby land and water are being used is an important starting point for identifying potentially exposed populations for a risk assessment. For complex facilities, especially those emitting PBT toxic air contaminants, it may be useful to follow DEQ's guidance on land and water use determinations (DEQ 1998a, DEQ 1998b). A specific combination of exposure locations, exposure routes, and land and water uses can be described as an exposure scenario. Once you have determined potential risks for the set of land and water use designations appropriate to the facility, any changes to designations means that risks should be re-evaluated in some manner. The key point is that if land and water uses change without a reassessment of risk appropriate for that site, the risk assessment may no longer be accurate.

## 2.4 Toxicity Assessment

The purpose of the toxicity assessment is to compile toxicity data for the toxic air contaminants a facility emits, and to estimate the relationship between the amount of exposure to a toxic air contaminant and the likelihood of adverse effects. You should evaluate the potential cumulative cancer risks and noncancer risks from all toxic air contaminants your facility emits. In most cases, facilities submitting a toxic air contaminant risk assessment can use the RBCs listed in OAR 340-245-8010 Table 2 to assess the toxicity of toxic air contaminants they emit, and will not need the additional information in this section. The RBCs provided in OAR 340-245-8010 Table 2 will be periodically updated, so facilities should verify they have the most recent version.

### 2.4.1 Noncancer Health Effects

The potential for noncancer health effects, such as organ damage, immunological effects, birth defects, or skin irritation, is assessed by comparison with what EPA calls a Reference Concentration (RfC) in units of  $\mu\text{g}/\text{m}^3$  or  $\text{mg}/\text{m}^3$ . The federal Agency for Toxic Substances and Disease Registry calls these concentrations Minimal Risk Levels (MRLs), and California's Office of Environmental Health Hazard Assessment calls them Risk Exposure Levels (RELs). For the purpose of this document, we will use the term noncancer toxicity reference value (TRV) for the selected RfC, MRL, or REL (see Appendix B). Cleaner Air Oregon rules require that risk assessments performed for the program use the TRVs provided in OAR 340-247-8010 Table 2.

A noncancer TRV is considered a threshold below which adverse effects are not likely even in sensitive groups. Often, TRVs are based on data from test animals. Because the goal of human health risk assessments is to protect humans, including sensitive humans, toxicologists use uncertainty factors to develop reference concentrations to ensure that the levels are protective of sensitive people.

Noncancer effects are evaluated by summing hazard quotients, as discussed in Section 3. The sum of hazard quotients for multiple toxic air contaminants is known as a hazard index (HI). HIs are most appropriately evaluated by individual target organ, although for simplicity, HIs are often calculated as a single total regardless of target organ. You can refine the HI evaluation by summing hazard quotients of toxic air contaminants with effects on the same organ system. Appendix F contains suggested tables of applicable organ systems for toxic air contaminants that will be acceptable to DEQ. Table F-1 is for chronic effects, and Table F-2 is for acute effects.

### 2.4.2 Cancer Effects

For cancer effects, the assumption is that there is no threshold for adverse effects. That is, we assume that exposure to even very small concentrations of the toxic air contaminant could contribute a small amount towards cancer risk. Because of this assumption, the toxicity of carcinogens is given not as a threshold concentration, but instead as a probability of getting cancer when exposed continuously to a concentration of  $1 \mu\text{g}/\text{m}^3$ . This value is called the inhalation unit risk (IUR) value, in units of risk per microgram per cubic meter  $(\mu\text{g}/\text{m}^3)^{-1}$ . For ease of use in assessing risk in the CAO program, IURs were converted to TRVs using a target excess cancer risk level of one in one million. Therefore, each TRV for cancer risk represents a one-in-one-million excess cancer risk.

### 2.4.3 Toxic Air Contaminant Classes

For some chemical classes, it is preferable to evaluate risk as a single value for the entire class

because the class exhibits toxicity by the same mechanism. Appendix E presents DEQ's recommendations on how to conduct evaluations for two important chemical classes:

- Chlorinated dibenzo-*p*-dioxins (CDDs), polychlorinated dibenzofurans (CDFs), and co-planar (dioxin-like) chlorinated biphenyl (PCB) congeners
- Polycyclic aromatic hydrocarbons (PAHs)

## 2.5 Risk-Based Concentrations

As discussed in Section 2.1, DEQ developed RBCs to simplify the risk assessment process. RBCs are available for the following exposure scenarios:

- Residential exposure, which includes long-term exposure to children and adults.
- Nonresident adult exposure, which includes workers in office buildings, commercial buildings, or industrial facilities.
- Nonresident child exposure, which includes schools and daycare facilities.
- Acute exposure, which includes areas where people may spend all or a portion of a day, such as parks, sports facilities, or agricultural fields.

TRVs serve as the basis for RBCs. To establish TRVs for each toxic air contaminant, DEQ relied on the scientific conclusions of agencies like the U.S. Environmental Protection Agency, and the Agency for Toxic Substances and Disease Registry. DEQ used the authoritative sources of chronic and acute TRVs identified in Appendix B.

Three adjustments of TRVs were made, if appropriate, to calculate RBCs. The first addresses scenario-specific consideration of exposure frequency and duration that are appropriate for chronic exposure scenarios. Another adjustment considers deposition and bioaccumulation of toxic air contaminants, which involves exposure other than by inhalation alone. This is a multipathway adjustment. The third adjustment is for early-life exposure to toxic air contaminants that exhibit greater toxicity to infants and children. These three types of adjustments are described in more detail in the sections below.

Adjustment factors for RBCs are provided in Table C-1. Appendix C shows how DEQ used the adjustment factors to develop the RBCs shown in OAR 340-245-8010 Table 2. Adjustment factors apply only to chronic exposure. None of the adjustment factors is appropriate or necessary for acute RBCs because of the short period of exposure being considered.

### 2.5.1 Exposure Frequency and Duration Considerations

Residential exposure assumes continual, long-term exposure. Because continual, long-term exposure is the basis of most chronic toxicity values, chronic TRVs are most directly applicable for residential exposure. For other types of exposure, including shorter term, nonresidential child exposure such as at schools, and worker exposure at commercial or industrial facilities, adjustments to TRVs are needed to take into consideration differences in exposure frequency and duration. Adjustment factors for chronic exposure are discussed in Appendix C.

## 2.5.2 Multipathway Adjustment Factors

If your facility emits persistent, bioaccumulative, and toxic (PBT) air contaminants, it is important to consider exposure through pathways other than air. For PBT toxic air contaminants, DEQ considered multipathway effects on residents in developing RBCs, which are used in all levels of risk assessment. DEQ developed multipathway adjustment factors (MPAFs) for residential exposure scenarios that consider:

- Inhalation of toxic air contaminants in air
- Deposition of airborne toxic air contaminants to backyard soil
- Contact with soil by incidental ingestion and dermal exposure
- Uptake into garden vegetables, and ingestion of vegetables, and
- Bioaccumulation into women, and infant ingestion of breastmilk

For nonresidential exposure, different MPAFs are used because some considerations, such as uptake into garden vegetables, are not appropriate. MPAFs do not include exposure scenarios that incorporate airborne deposition of toxic air contaminants to:

- Agricultural land
- Livestock grazing areas
- Drinking water reservoirs
- Bodies of water used for fishing

If PBT toxic air contaminant emissions from your facility could impact the above areas, DEQ may require a more complex evaluation of risk considering multipathway exposure, even if emissions screen out at the Level 1, 2, or 3 risk assessments using default MPAFs [OAR 340-245-0050(12)]. In practice, the list of PBTs can be considered those listed in Table C-1 for which there are MPAFs, because for other PBTs, insufficient information is available to quantitatively evaluate risk. The list of PBTs may be expanded in the future as additional information becomes available.

DEQ will consider toxic air contaminant deposition rates, and may consider other factors such as size of impacted area and degree of use. DEQ recommends that the process for calculating risk from PBT toxic air contaminants start with development of a conceptual site model more extensive than the default assumptions. This additional multipathway risk evaluation may sufficiently address DEQ's concerns without requiring the need for a Level 4 risk assessment.

## 2.5.3 Early-Life Exposure Adjustment Factors

Carcinogens that act by a mutagenic mode of action can have greater toxicity during early-life stages (EPA 2005a). In these cases, we need to adjust the cancer TRV. Currently, the toxic air contaminants of primary interest for consideration of early-life exposure are listed in Table C-1 with early-life adjustment factors (ELAFs). Appendix D shows the derivation of ELAFs. As more information becomes available, EPA may determine that additional carcinogens act by a mutagenic mode of action. If this occurs, DEQ may undertake rulemaking to expand the list of toxic air contaminants for which ELAF values are needed, and revise RBCs accordingly.

## 2.6 Exposure Assessment

Estimation of exposure involves the identification of exposure pathways, scenarios, and routes. The initial identification of these elements is in the conceptual site model, which should be included in the

modeling protocol. An exposure pathway is the course a toxic air contaminant takes from a source to an exposed organism (EPA 1989). Exposure scenarios are comprised of one or more exposure routes appropriate to the potentially exposed population. An exposure route is the way a toxic air contaminant comes in contact with a person. Inhalation is the primary exposure route for air emissions, although other routes (ingestion, dermal contact) may be important for PBT toxic air contaminants.

For CAO, chronic exposure scenarios are residential, non-residential adult (worker), and non-residential child (schools and daycare facilities). Acute risk should be evaluated for all the chronic exposure locations. In determining whether it is appropriate to evaluate acute risk at additional locations, consider the following: Is it reasonable for people to spend two or more hours in a day at this location? All parks, sports facilities, and agricultural fields are likely relevant locations for acute exposure. It is not necessary to evaluate acute exposure to transient situations, such as walking on a commercial sidewalk, hiking on a trail, driving on a road, or paddling a river. However, if people are known or expected to stay in an area, you should evaluate acute exposure. Examples include picnicking or camping in parks or along hiking trails, or fishing from a riverbank, or from a boat in the river. Typically people will not spend more than two hours at a small cemetery. However, if larger cemeteries have caretakers or other workers, acute risks should be evaluated (in addition to worker exposure).

Level 1, 2, 3 and 4 risk assessments include evaluating potential exposures for all relevant exposure scenarios through some form of air dispersion modeling, from lookup tables to screening or refined models.

When completing a Level 4 risk assessment for a source that emits PBT toxic air contaminants, additional scenarios such as agricultural or recreational use may be relevant. Details about how to evaluate these exposure scenarios are not provided in this document, and we recommend discussing them with DEQ prior to submitting a modeling protocol and a risk assessment work plan.

### **2.6.1 Air Dispersion Modeling**

This section provides a brief overview of air dispersion modeling and its relationship to risk assessments. More complete recommended protocols for modeling are found in DEQ's *Recommended Procedures for Air Dispersion Modeling* (DEQ 2022).

A primary element of an exposure assessment for toxic air contaminants is air quality dispersion modeling, which results in exposure concentrations at all levels of the risk assessment, from Level 1 through Level 4. These exposure concentrations, and their comparison to the RBCs, are the foundation of the risk assessment. Essentially, a dispersion model is a mathematical approximation of the physical, and sometimes chemical, processes in the atmosphere that disperse emissions, and calculates air concentrations of toxic air contaminants. Dispersion modeling estimates these concentrations at specific geographic points called modeling receptors. Modeling receptors can be positioned to coincide with exposure locations, for example residential or worker areas. Modeling receptors are typically arrayed in a grid, although they can also be positioned at specific locations, such as houses or schools. Modeling receptors are virtual points in space and should not be confused with the term "receptor", commonly used to identify humans in risk assessments.

There are four risk assessment levels, and the role of dispersion modeling for each level is briefly described below.

#### **Level 1**

For Level 1, DEQ developed lookup tables of dispersion factors (OAR 340-245-8010 Table 3) that are based on pre-run modeling results, so it is not necessary for facilities to conduct modeling. Facilities can calculate concentrations for comparison to the RBCs to assess risk using emissions and dispersion factors obtained from the tables.

## **Level 2**

Risk assessments at Level 2 require the direct use of the dispersion model AERMOD, or another approved model. AERMOD is an EPA approved dispersion model for regulatory modeling, and is the primary refined dispersion model for the Cleaner Air Oregon program. AERMOD-MAKEMET and AERSCREEN are screening versions of AERMOD using worst case screening meteorology. There are distinct advantages to using AERMOD-MAKEMET, including the ability to model multiple emission points in a single model run. AERSCREEN requires separate runs for each emission unit, and may be more conservative since the individual maximum concentrations from single runs must be added to get a total concentration. In addition, if after the Level 2 analysis a Level 3 analysis is preferred, the AERMOD-MAKEMET model input file can be re-used for a full AERMOD analysis with the replacement of MAKEMET data with actual meteorology. For these reasons DEQ recommends AERMOD-MAKEMET for the Level 2 analysis.

## **Levels 3 and 4**

For risk assessment Levels 3 and 4, the full AERMOD model should be used with actual representative meteorological data and a gridded array or field of modeling receptors where concentrations will be evaluated.

## **Modeling Protocol and Reporting**

Prior to submitting a risk assessment, the owner or operator must prepare and get DEQ approval of a modeling protocol. Detailed information on developing a modeling protocol and for using AERMOD-MAKEMET, AERSCREEN, or AERMOD, and their pre- and post-processors, is in DEQ's *Recommended Procedures for Air Dispersion Modeling* (DEQ 2022). After the modeling protocol is approved and implemented, a report must be prepared providing the results of the modeling and risk assessment. Because air dispersion modeling is the exposure assessment element of the risk assessment, the modeling results are included in the risk assessment report for all levels. The amount of detail required in the modeling portion of the risk assessment report varies by the risk assessment level selected.

## **2.6.2 Use of Air Monitoring Data in Risk Assessments**

You may request to conduct ambient air monitoring to supplement air modeling after completion of an approved Level 3 or Level 4 risk assessment and Risk Reduction Plan (if applicable). There are a number of complexities to using air monitoring data in a risk assessment. The presence of multiple sources of toxic air contaminants near the facility can complicate ambient monitoring results. This requires simultaneous monitoring upwind and downwind of a facility. This is further complicated by varying wind directions over the year. A year of monitoring results may reasonably provide an annual average concentration at the monitoring station, suitable for comparison with chronic RBCs; however, it is far more difficult to determine the highest daily concentration that could occur at a monitoring location. This uncertainty could underestimate acute risks.

Ambient monitoring would likely take a minimum of 1.5 years to: 1) develop an adequate monitoring protocol, 2) receive DEQ approval for monitoring, and 3) obtain and deploy sampling equipment. Another six months may be required to analyze the data, develop conclusions, and obtain DEQ approval of the conclusions in a final, revised risk assessment based on the monitoring data. The monitoring protocol should include data quality objectives, and describe how exposure concentrations will be used to evaluate risk in a revised risk assessment.

In consideration of the above complexities, if DEQ approves monitoring results, the monitoring results will be used to update the total risk at a facility and compare with Risk Action Levels. Sometimes ambient air monitoring may result in non-detect values for toxic air contaminants. Non-detect values may be handled according to the approach presented in Appendix G.

## 2.7 Risk Characterization

Comparing modeled or measured exposure concentrations with RBCs is an efficient means of determining risk at a facility, and is an integral part of the CAO approach. In general, the risk for an exposure scenario is:

Equation 2.1

$$\text{Risk} = \text{Concentration} / \text{RBC}$$

Where:

Risk = excess cancer risk, chronic hazard index, or acute hazard index

Concentration = modeled or measured exposure concentration ( $\mu\text{g}/\text{m}^3$ )

RBC = risk-based concentration ( $\mu\text{g}/\text{m}^3$ ) appropriate for the exposure scenario

Details on how to perform the calculations are provided in Section 3.1.2. Appendix A contains example tables of how to present the results of the risk calculations, and how to compare final facility-wide risk results with risk action levels.

## 2.8 Uncertainty Evaluation

CAO rules require that a quantitative or qualitative uncertainty evaluation be included in Level 3 and Level 4 risk assessments. Often it is difficult to quantitatively evaluate uncertainty. Generally, though, you can make a determination that an uncertainty will likely result in an underestimate or overestimate of risk. In some cases it will be unknown whether risks will be under- or over-estimated. There are various types of uncertainties associated with a risk assessment, including the following four major categories:

- Selection of toxic air contaminants for evaluation
- Emission rate calculations
- Exposure assessment assumptions
- Derivation of toxicity values

If there are no RBCs for some of the emitted chemicals at a facility, these chemicals cannot be quantitatively evaluated, which will result in an underestimation of risk. Also, a facility may not be aware of all chemicals emitted by some processes.

Often emission rates are calculated using published emission factors developed by EPA or other entities. In general, these factors are designed to be protective, and should not underestimate emissions and risk. If these factors are applied to similar processes for which the factors were not developed, emission estimates may be under- or over-estimated. In some cases the emission factor data is sufficiently robust to include statistical information, including mean and median values, as well as a range. Uncertainty in emission rates estimates can be partially quantified by evaluating the range of emission factors that could be used.

One approach to substantially reduce uncertainty in emission estimates is to conduct source testing. Collecting actual emission data will reduce uncertainty related to emissions and associated risk.

Uncertainty associated with exposure can include potential inaccuracies in the emission inventory,



variability in estimates of emission rates, uncertainties in air dispersion models, and protectiveness inherent in the exposure assumptions incorporated into the derivation of RBCs. Toxic air contaminants that are missed in the emission inventory, or have underestimated emission rates, will result in an underestimation of risk. Protective assumptions used in models will likely overestimate risk. However, in cases with close proximity of receptors to the source, risk may be underestimated.

There is often high uncertainty associated with monitoring air concentrations. To reduce uncertainty, a sufficient number of monitors need to be deployed, and they need to be appropriately placed to obtain representative data. Detection limits need to be adequate to detect toxic air contaminants above RBCs. Uncertainty associated with evaluating non-detected concentrations needs to be discussed.

Sources of uncertainty for toxicity values can be discussed. For carcinogens, EPA has weight of evidence categories that can be presented and discussed. For noncarcinogens, TRVs have associated uncertainty factors that can also be presented and discussed. Almost all risk assessments are expected to include a summation of risk from multiple toxic air contaminants. The assumption of dose additivity (inherent in the rule requirements for summing risk) does not consider possible synergistic or antagonistic effects. The potential for under- or over-estimating summed risks can be discussed. Evaluating noncancer risk by target organ (Appendix F) is one way to reduce uncertainty in the noncancer risk evaluation.

In some regulatory programs, such as DEQ's Cleanup Program, probabilistic analyses can be performed. A probabilistic risk assessment inherently quantifies uncertainty. However, a probabilistic approach such as a Monte Carlo analysis is not appropriate to evaluate uncertainty for a facility in the Cleaner Air Oregon program. This is because, unlike Cleanup Program rules, the CAO statute and rules do not contain provisions for determining acceptable risk based on probabilistic results. Risk action levels in the CAO program are for deterministic results.

# 3. CONDUCTING LEVEL 1 THROUGH LEVEL 4 RISK ASSESSMENTS

## 3.1 Introduction

To conduct a risk assessment, you need to calculate concentrations of toxic air contaminants at the exposure locations identified in the conceptual site model. You may assess risk under the CAO rules without conducting dispersion modeling for emissions from your facility by using the Level 1 Risk Assessment Tool of dispersion factors (Section 3.2). If you are unable to use a Level 1 approach, or want to have a more refined assessment, then you may complete an assessment using the Level 2 simple model to calculate toxic air contaminant concentrations (Section 3.3).

Level 1 and Level 2 risk assessments include conservative assumptions that are very likely to overestimate risk. If you want to further refine your risk assessment and eliminate these conservative assumptions, you may conduct more detailed site-specific air dispersion modeling (Levels 3 and 4). This will allow you to more accurately quantify the risks posed by toxic air contaminants at your facility in order to create a more site-specific evaluation of potential exposure.

The key element of a Level 3 or Level 4 risk assessment is the use of more sophisticated air dispersion modeling. This is discussed in Section 3.4. If you decide to undertake a more detailed evaluation beyond air dispersion modeling, or if DEQ requires consideration of non-inhalation pathways that are important and not covered by the assumptions incorporated into RBCs, a Level 4 risk assessment will be required (Section 3.5). You are required to prepare a risk assessment work plan and modeling protocol for either a Level 3 or 4 risk assessment. The results of Level 3 and Level 4 risk assessments are estimates of risk, not measurements of actual risk.

Figure 1 provides an overview of the exposure modeling and risk assessment process.

### 3.1.1 Emission Rate Determination

A common element in all levels of risk assessment is the estimation of emission rates for toxic air contaminants identified on the emission inventory. DEQ has a form available for completing the emission inventory, with instructions to assist facilities. For the emission rates at any level of risk assessment, select from among the following:

- All existing sources may evaluate risk based on actual emissions from the previous year. However, if you want to be permitted at either capacity, or a requested Potential to Emit (PTE), that is different from your actual emissions, you must assess risk at this requested level of emissions as well.
- Requested PTE activity levels must be included, along with your actual activity levels, in your submitted emission inventory, as noted in OAR 340-245-0040(4)(a)(B) and (4)(b)(B).
- In order to evaluate if you are a *de minimis* source, you must assess toxic air contaminant emissions at the capacity to emit, as defined in OAR 340-200-0020(14).

Note that DEQ will evaluate and approve each facility's PTE when reviewing emission inventories. A source has the choice to have DEQ set a Source Risk Limit based on the source's actual emissions, existing PTE used to calculate Plant Site Emission Limits, or a requested PTE (as noted above) for toxic air contaminants.

It is important to clearly identify the form of the chemical being emitted when modeling for the risk assessment, because there can be large differences in the toxicity of different forms of chemicals. For example, this may require accounting for different oxidation states of a metal species like chromium, distinguishing between trivalent and hexavalent forms. This is also discussed in Section 4.2.2.

### 3.1.2 Comparison of Exposure Concentrations with RBCs

Another common element in all levels of risk assessment is the comparison of modeled or measured air concentrations at exposure locations with their corresponding RBCs. How air concentrations are estimated varies by the level of risk assessment, as discussed below for Levels 1 through 4 in Sections 3.2 to 3.5.

The evaluation should consider total toxic air contaminant emissions from all non-exempt toxic emission units (TEUs) at the facility. In Level 1 (using the lookup table) and Level 2 (using simple modeling), exposure concentrations will be calculated for each TEU. You need to sum these concentrations to get the total risk from the facility. In Levels 3 and 4, the more complex modeling evaluates emissions from all TEUs at the same time, without the need for later summation of calculated concentrations. Exposure concentrations need to be estimated for all applicable exposure locations (residential, commercial/industrial, school/daycare, and acute).

Once an exposure concentration for each chemical is estimated at all relevant exposure locations, divide the maximum calculated exposure concentration by the respective RBC for that toxic air contaminant from OAR 340-245-8010 Table 2. This will determine the potential risk for that single toxic air contaminant emitted from a single TEU. Sum the results for all the relevant toxic air contaminants, and then sum the results for all the different TEUs to get a facility risk. You should calculate a facility risk for each of the different exposure scenarios and the corresponding RBCs (chronic cancer, chronic noncancer, and acute noncancer).

In summary, for each exposure location, calculate the following:

Equation 3.1

$$\text{Excess Cancer Risk} = \sum_{j=1}^{TEU\ m} \sum_{i=1}^{chemical\ n} \frac{ltConc_{ij}}{caRBC_i}$$

Equation 3.2

$$\text{Chronic Hazard Index} = \sum_{j=1}^{TEU\ m} \sum_{i=1}^{chemical\ n} \frac{ltConc_{ij}}{ncRBC_i}$$

Equation 3.3

$$\text{Acute Hazard Index} = \sum_{j=1}^{TEU\ m} \sum_{i=1}^{chemical\ n} \frac{stConc_{ij}}{acuteRBC_i}$$

Where:

Excess cancer risk = probability of developing cancer (above background rate), expressed as a per million rate

Chronic hazard index = sum of hazard quotients used to evaluate chronic noncancer health risk (can be calculated separately by target organ)

Acute hazard index = sum of hazard quotients used to evaluate acute noncancer health risk (can be calculated separately by target organ)

ltConc<sub>ij</sub> = Calculated long-term (annual) concentration of toxic air contaminant *i* from emission source *j*

stConc<sub>ij</sub> = Calculated short-term (daily) concentration of toxic air contaminant *i* from emission source *j*

caRBC<sub>i</sub> = Cancer Risk-Based Concentration for toxic air contaminant *i*

ncRBC<sub>i</sub> = Noncancer Risk-Based Concentration for toxic air contaminant *i*

acuteRBC<sub>i</sub> = Acute noncancer Risk-Based Concentration for toxic air contaminant *i*

Finally, compare these calculated total risk values with the RALs in OAR 340-245-8010 Table 1.

### 3.1.2.1 Consideration of Adjusted Hazard Index RAL

Under Senate Bill 1541, DEQ could consider and propose that the Environmental Quality Commission (EQC) by rule adjust the TBACT RAL at existing sources to a hazard index value other than 5 for certain non-cancer toxic air contaminants. To be eligible for this adjustment, a chemical must be expected to have developmental human health effects associated with pre- or post-natal exposure, or other severe human health effects. SB 1541 specified that any adjusted RAL may be no less than a hazard index of 3. DEQ established an advisory committee in 2018, and considered their recommendations prior to developing a list of toxic air contaminants for which the TBACT RAL for existing sources would be an HI of 3 rather than 5. These chemicals, with developmental or other severe health effects, are identified in Appendix F along with target organs. The statutory requirement limiting a noncancer RAL to no less than 3 sunsets on January 1, 2029, after which the EQC may establish new noncancer RALs.

Having two TBACT hazard index RALs for existing sources requires an additional step in evaluating compliance with the RALs. Revised rules (OAR 340-245-0200) specify the use of a Risk Determination Ratio (RDR) approach. In this approach, risk is first calculated similar to the calculation of risk presented in Equations 3.2 to 3.3. Chemicals designated as HI3 are compared with a TBACT RAL hazard index of 3 to calculate risk from HI3 chemicals. Chemicals designated as HI5 are compared with a TBACT level RAL hazard index of 5 to calculate risk from HI5 chemicals. To fully consider cumulative effects, the risk from HI3 chemicals and the risk from HI5 chemicals are added together to produce a Risk Determination Ratio (RDR). If the resulting RDR is greater than 1.0, the TBACT level RAL for an existing facility is exceeded. The RDR for the Risk Reduction Risk Action Level (RAL) is 2.0, and the RDR for the Immediate Curtailment RAL is 4.0 (OAR 340-245-8010 Table 1).

This approach can be expressed as follows:

Equation 3.4

$$Risk_{HI3} = \sum_{HI3 \text{ chemicals}} \frac{\text{Concentration}}{RBC}$$

Equation 3.5

$$Risk_{HI5} = \sum_{HI5 \text{ chemicals}} \frac{\text{Concentration}}{RBC}$$

Equation 3.6

$$Risk \text{ Determination Ratio} = \frac{Risk_{HI3}}{3} + \frac{Risk_{HI5}}{5}$$

Where:

HI3 = Toxic air contaminants assigned noncancer TBACT Risk Action Level of 3 (OAR 340-247-8010, Table 2 and OAR 340-245-8010, Table 2).

HI5 = Toxic air contaminants assigned noncancer TBACT Risk Action Level of 5 (OAR 340-247-8010, Table 2 and OAR 340-245-8010, Table 2).

Concentration = monitored or modeled concentrations of toxic air contaminant at exposure location for use in risk assessment.

RBC = risk-based concentrations in OAR 340-245-8010 Table 2.

The final RDR is expressed to one decimal place. Appendix A shows examples of how to evaluate noncancer risk at facilities with both HI3 and HI5 chemicals.

### **Risk Action Levels**

RALs are levels of risk that prescribe the actions a regulated facility is required by DEQ to undertake under Cleaner Air Oregon rules. As risks from the source increase, so do the more corrective actions the facility is required to take, as specified by the applicable RAL.

The RALs reflect the challenges existing facilities could face in retrofitting existing equipment to meet new lower risk requirements under the Cleaner Air Oregon program. Cleaner Air Oregon rules include higher RALs for existing facilities than for new or reconstructed facilities, which are easier to design with risk reduction in mind.

In recommending RALs to the Environmental Quality Commission (EQC), DEQ considered risk levels used in decision making by federal agencies like the EPA and the Agency for Toxic Substances and Disease Registry. DEQ also considered RALs used by other states that already have health risk-based industrial air toxics programs. In addition, DEQ considered the overall non-industrial level of risk from air toxics in Oregon. Finally, DEQ considered risk benchmarks set in Oregon statute following passage of Senate Bill 1541 in the 2018 Oregon Legislature.

Cleaner Air Oregon considers risk for three categories: chronic cancer, chronic noncancer, and acute noncancer. Risk from toxic air contaminants that contribute to each of those categories of risk are calculated and compared against the appropriate RAL separately. Cancer and noncancer risks are calculated and expressed differently. The RALs for Cleaner Air Oregon reflect those differences in that there are separate sets of RALs for cancer and noncancer risks. RALs are shown in OAR 340-245-8010 Table 1.

Senate Bill 1541 included a January 1, 2029 sunset provision for its requirements that set minimum levels for the existing source TBACT (toxics best available control technology) level and Risk Reduction Level. After that date, the Environmental Quality Commission may establish new risk levels for those RALs through rulemaking. The TBACT RALs could be set at values not lower than 25 in one million excess cancer risk and a hazard index value of 1.

### 3.1.3 Emission Modeling Approaches for Risk Calculations

This section outlines the different ways a facility may model emissions of toxic air contaminants at the facility. Figure 3 summarizes four approaches. DEQ considers any of these approaches acceptable for meeting the requirements of CAO rules. Appendix A provides example calculations for the different approaches. The approaches are also presented in DEQ's *Recommended Procedures for Air Dispersion Modeling* (DEQ 2022).

#### Approach A: Unadjusted Emission Rate

The standard approach for calculating risk is to:

- Develop a toxic air contaminant emission rate. The unadjusted rate can be the actual rate based on emissions from the previous year (for existing facilities), or PTE.
- Apply a dispersion factor from a lookup table, or use a computer model to calculate an air concentration, and
- Divide the concentration by an RBC to calculate risk.

This is shown by the following sets of equations:

##### Level 1 Lookup Table

Equation 3.7

$$ER \text{ (lb/yr)} \times DF \text{ (}\mu\text{g/m}^3 \text{ per lb/yr)} = C_{\text{air}} \text{ (}\mu\text{g/m}^3\text{)}$$

Equation 3.8

$$C_{\text{air}} \text{ (}\mu\text{g/m}^3\text{)} / \text{RBC (}\mu\text{g/m}^3\text{)} = \text{Risk (per million cancer risk or hazard quotient)}$$

##### Level 2, 3, or 4 Modeling

Equation 3.9

$$ER \text{ (g/s)} \rightarrow \text{AERMOD} \rightarrow C_{\text{air}} \text{ (}\mu\text{g/m}^3\text{)}$$

Equation 3.10

$$C_{\text{air}} \text{ (}\mu\text{g/m}^3\text{)} / \text{RBC (}\mu\text{g/m}^3\text{)} = \text{Risk (per million cancer risk or hazard quotient)}$$

[Post modeling step]

Where:

ER = emission rate (lb/yr, lb/day, or g/s)

DF = dispersion factor ( $\mu\text{g/m}^3$  per lb/yr, or  $\mu\text{g/m}^3$  per lb/day)

$C_{\text{air}}$  = air concentration ( $\mu\text{g/m}^3$ )

RBC = risk-based concentration ( $\mu\text{g/m}^3$ )

If unadjusted emission rates are used in AERMOD or another air dispersion model, the resulting concentration output in  $\mu\text{g/m}^3$  is available for each toxic air contaminant. Using unadjusted emission rates will provide calculated exposure concentrations, not risk. To calculate risk, an additional calculation of dividing the concentration by the appropriate RBC is needed for each toxic air contaminant at each appropriate exposure location (residential, non-residential child, non-residential adult, and acute).

## Approach B: Unit Emission Rate

For convenience in modeling, often a unit emission rate of 1 gram/second (g/s) is used. The resulting air concentration must be converted for each toxic air contaminant for a given emission unit. These calculations need to be performed in a spreadsheet or database after modeling. Similar to using unadjusted emission rates, using a unit emission rate with post-processing will provide calculated exposure concentrations, not risk. To calculate risk, dividing the concentration by the appropriate RBC is needed for each toxic air contaminant at each appropriate exposure location (residential, non-residential child, non-residential adult, and acute).

## Approach C: Risk-Equivalent Emission Rate

An alternative method that could reduce the post-modeling calculation effort uses a Risk-Equivalent Emission Rate (REER), instead of an unadjusted emission rate or unit emission rate. Approach C for calculating risk is to:

- Develop a toxic air contaminant emission rate
- Divide the emission rate by an RBC to calculate a REER, and
- Apply a dispersion factor from a lookup table, or use a computer model to calculate risk

This is shown by the following sets of equations:

### Level 1 Lookup Table

Equation 3.11

$$[ER \text{ (g/s)} / RBC \text{ (}\mu\text{g/m}^3\text{)}] \times DF \text{ (}\mu\text{g/m}^3 \text{ per g/s)} = \text{Risk (per million cancer risk or hazard quotient)}$$

Because no modeling is required for a Level 1 analysis, this process is no different from Approach A.

### Level 2, 3, or 4 Modeling

To implement this alternative approach with AERMOD dispersion modeling in Levels 2, 3, and 4, emission rates for each toxic air contaminant at each emission unit are normalized to risk by dividing by the appropriate RBC. This calculation needs to be performed for each toxic air contaminant, at each emission unit, for all relevant exposure locations (RBCs). The result is a REER value in units of g/s per  $\mu\text{g/m}^3$ .

Equation 3.12

$$ER \text{ (g/s)} / RBC \text{ (}\mu\text{g/m}^3\text{)} = REER \text{ (g/s per } \mu\text{g/m}^3\text{)}$$

[Pre-modeling step]

Equation 3.13

$$REER \text{ (g/s per } \mu\text{g/m}^3\text{)} \rightarrow \text{AERMOD} \rightarrow \text{Risk (per million cancer risk or hazard quotient)}$$

The calculation of a REER normalizes the emission rate to risk, either an excess cancer risk of one in one million for carcinogens, or a hazard quotient of 1 for noncarcinogens. Because REER is directly proportional to risk, REERs for the various toxic air contaminants can be added together at each TEU. After running a model such as AERMOD, the concentration results reported as  $\mu\text{g/m}^3$  are now equivalent to units of risk (risk per million for carcinogens, and hazard index for noncarcinogens). This approach of modeling risk substantially minimizes the time and effort for post-processing calculations

necessary using an approach that models chemical emission rates. DEQ considers this method mathematically equivalent to performing separate risk calculations after modeling.

#### **Approach D: Unit Emission Rate with Risk-Equivalent Emission Rate**

There are advantages for using a unit emission rate approach, and to using a risk-equivalent emission rate approach. In many cases, it may be advantageous to use a combination of both approaches. With this hybrid approach, air dispersion modeling is performed using a unit emission rate. In a separate calculation performed in a spreadsheet, unadjusted toxic air contaminant emission rates are divided by RBCs to calculate a REER. The REER from this pre-processing step can then be multiplied by the calculated air concentration using the unit emission rate to calculate total risk (risk per million for carcinogens, and hazard index for noncarcinogens). Approach D reduces post-processing efforts while maintaining the flexibility of using unit emission rates.

## **3.2 Modeling Protocol and Risk Assessment Work Plan**

As discussed in Section 2.6.1, prior to submitting the risk assessment, the owner or operator must prepare a modeling protocol. This applies to modeling at all levels of risk assessment. In addition, you are required to prepare a risk assessment work plan for Level 3 and Level 4 risk assessments. Because the modeling protocol covers the exposure assessment portion of the risk assessment, the risk assessment work plan can include the modeling protocol to avoid duplication of efforts. Details on the modeling protocol and report are included in DEQ's modeling procedures document (DEQ 2022).

The modeling protocol should include a map with topographic features, a facility plot plan with site features identified, and maps showing exposure locations and modeling receptors. Modeling receptor locations for CAO should extend from no less than 2 km and up to 10 km from the facility, but must include all areas where modeled risk is at or above 0.5 in 1 million excess cancer risk, or at a hazard index of 0.5 for chronic and acute noncancer risk. For point, area, and volume sources, provide details relevant to modeling emissions, such as location and dimensions. Specify which approach (A, B, C, or D) discussed in Section 3.1.3 will be used.

The risk assessment level selected will affect the information submitted in the modeling protocol, as shown in Figure 1. Submit the following information in the modeling protocol (and risk assessment work plan, if relevant):

- Toxic Air Contaminants (TACs), with and without RBCs
  - The total annual and acute emissions from the facility of each TAC.
  - The RBC for each TAC, from OAR 340-245-8010 Table 2.
- Toxic Emission Units (TEU) and their respective TACs – Level 1
  - Location of each TEU in a figure.
  - Emission type (point or fugitive) for each TEU.
  - Stack height for point sources.
  - Building dimensions for fugitive sources.
  - Annual and acute emission rates, in lbs/yr or lbs/day, respectively, of each TAC by TEU.
- Toxic Emission Units (TEU) and their respective TACs – Levels 2-4
  - Location of each TEU.
  - Emission type (point, area, volume, etc.) for each TEU.
  - The model-ready stack parameters for each TEU.
  - Annual and acute emission rates, in g/s, of each TAC by TEU



- Exposure Locations
  - Levels 1, and 2 (AERSCREEN), distances to all exposure locations should be identified
  - Levels 2 (AERMOD-MAKEMET), 3, and 4, define a receptor grid and identify exposure locations
  - Zoning maps and a crosswalk between modeling receptors and exposure location assignment
- Meteorological Dataset
  - Level 1, no need to develop meteorology
  - Level 2, worst case meteorological dataset (MAKEMET) is required
  - Levels 3-4, meteorological data for input to AERMOD should be representative of the facility location

### 3.3 Level 1 Risk Assessment

The Level 1 approach allows you to use toxic air contaminant emission rates from your facility to calculate exposure concentrations. If you use a Level 1 risk assessment, a simple modeling protocol, including methods to combine stacks and basis for distance to exposure locations, must be approved by DEQ before the Level 1 risk assessment is performed.

To perform a Level 1 risk assessment, use the dispersion factors listed in OAR 340-245-8010 Table 3. If your emissions come from a stack, use Table 3A to estimate chronic risk and Table 3B to estimate acute risk. If you have volume fugitive emissions that do not come from a stack, use Table 3C to estimate chronic risk, and Table 3D to estimate acute risk. Because DEQ has done the dispersion modeling to develop these dispersion factors, it is not necessary for you to run an air dispersion model. The table is designed for both point source emissions from discrete stacks that have quantifiable dimensions, and volume fugitive emissions that do not have a discrete emission point, including emissions from building doors and windows, or from areas where paint, solvent, or other emissions are generated. These factors are shown in OAR 340-245-8010 Tables 3A (annual exposure) and 3B (24-hour exposure) for stack emissions, and Tables 3C (annual exposure) and 3D (24-hour exposure) for fugitive emissions.

OAR 340-245-8010 Table 3 may not be used if there is elevated terrain higher than the stack height within 1.5 kilometers from the source. In this case, the assumptions used to develop the dispersion factors in Table 3 are not valid for the source, and Level 2 or Level 3 modeling will be necessary. If the stack height is less than 5 meters, or the closest exposure location is less than 50 meters away, consult with DEQ to determine the applicability of using a Level 1 risk assessment.

For stack emissions, the use of OAR 340-245-8010 Tables 3(A,B) requires the stack height and distance from the stack to the nearest exposure locations. For volume fugitive emissions, Tables 3(C,D) require the building area and height, and distances from the building to the nearest exposure locations. For sources that do not readily fit the scenarios presented in Table 3, refer to the *Recommended Procedures for Air Dispersion Modeling* (DEQ 2022). For example, a wastewater treatment plant may have ponds that emit toxic air contaminants that would be evaluated as area fugitive sources using Level 2 or Level 3 air dispersion modeling.

You will likely have four types of exposure locations to evaluate (residential, commercial/industrial, school/daycare, and acute), as discussed in Section 2.6. Determine the closest distance to all the relevant exposure locations. Find the dispersion factor in the table for the exposure location distance and a given stack height for stacks, or building height and dimensions for volume fugitive emissions. For each exposure location, multiply the emission rate from a facility stack or fugitive emitting process

by the dispersion factor.

Stack and volume fugitive emission rates must be in the same units as the table, such as pounds/day for toxic air contaminants with acute effects, and pounds/year for those with chronic effects. The result of the calculation will be air concentrations in units of micrograms per cubic meter,  $\mu\text{g}/\text{m}^3$ .

Next, for each exposure location, divide the calculated exposure concentration by the respective RBC for that toxic air contaminant from OAR 340-245-8010 Table 2, and sum the results as discussed in Section 3.1.2. Compare these calculated total risk values with the RALs in OAR 340-245-8010 Table 1.

### **DEQ's Development of Dispersion Factors in OAR 340-245-8010 Table 3**

To generate the dispersion factors provided in OAR 340-245-8010 Table 3, DEQ first developed a series of reasonable maximum assumptions associated with stack height, such as stack diameter, stack flow rate, and building dimensions. We used meteorology data from six airport sites representing different regions of the state (Portland, Salem, Eugene, Medford, Redmond, and Hermiston). We then used AERMOD to estimate air concentrations at distances from 50 meters to 1,000 meters from the stack. We averaged the results for each of the exposure location distances from the six meteorological sites to develop dispersion factors in units of  $\mu\text{g}/\text{m}^3$  per pounds/year for chronic exposure. For acute exposure, we used the maximum result at each exposure location distance to develop dispersion factors in units of  $\mu\text{g}/\text{m}^3$  per pounds/day. The results are shown in OAR 340-245-8010 Tables 3A (annual exposure) and 3B (24-hour exposure).

DEQ developed fugitive emission dispersion factors in a similar fashion using a series of reasonable maximum assumptions associated with building area and height, and a single set of meteorology data that was a conservative representation of areas in the state. As with stack emissions, we used AERMOD to estimate air concentrations at distances from 50 meters to 1,000 meters from the building. The results were used to develop tables of dispersion factors in  $\mu\text{g}/\text{m}^3$  per pounds/year for chronic exposures (OAR 340-245-8010 Table 3C), and  $\mu\text{g}/\text{m}^3$  per pounds/day for acute exposures (OAR 340-245-8010 Table 3D)

Example 1 in Appendix A shows a simple Level 1 risk assessment.

The Level 1 risk assessment process is meant to be conservatively protective, and was designed primarily to assist smaller facilities in their risk assessments. It is likely that for larger, more complex facilities, a Level 1 analysis may overestimate risk. DEQ does not expect any health concerns at facilities that use a Level 1 analysis and calculate values below Source Permit Level RALs. Because risk estimated using a Level 1 risk assessment is generally much higher than actual risk, higher results obtained using this method may indicate a need for further evaluation. Completing a Level 2 or 3 risk assessment will likely show lower, more accurate estimates of risk.

For a Level 1 assessment, you can treat multiple stacks at your facility in one of two ways:

- Add the toxic air contaminant concentrations in  $\mu\text{g}/\text{m}^3$  from all individual stacks to estimate an aggregate concentration obtained using the procedures above and then compare the concentration to the RBC for that toxic air contaminant; or
- Group the stacks and their emissions into a single stack, and use the information in OAR 340-245-8010 Table 3 to determine a dispersion factor to apply to the grouped emissions in order to estimate an air concentration for comparison to the RBC for that toxic air contaminant.

Similarly, you can address volume fugitive emissions from multiple buildings by either adding toxic air contaminant air concentrations to estimate an aggregate concentration, or grouping emissions into a single building. DEQ can assist with information about methods to group stacks and buildings.

For a stack height between the values shown in OAR 340-245-8010 Tables 3A and 3B, you may either use the next lowest stack height, or interpolate the dispersion factor. Similarly, for an exposure location distance between values shown in the table, you may either use the next lower distance, or interpolate the dispersion factor. For stack heights greater than 50 meters, use the appropriate dispersion factor for 50 meters. For exposure locations greater than 1,000 meters from your facility, use the appropriate dispersion factor at 1,000 meters.

Obtain stack heights and distances to exposure locations for your facility. However, in the absence of a known stack height and exposure location distance, you may use the annual dispersion factor ( $0.0033 \mu\text{g}/\text{m}^3$  / pounds/year) and daily dispersion factor ( $8.3 \mu\text{g}/\text{m}^3$  / pounds/day) for a stack height of 5 meters and an exposure location distance of 50 meters.

In evaluating volume fugitive emissions, for an exposure location distance between the values shown in the OAR 240-245-8010 Tables 3C and 3D, you may either use the next lowest distance, or interpolate the dispersion factor. For exposure locations greater than 1,000 meters from the building, you may use the appropriate dispersion factor at 1,000 meters. In the absence of known building dimensions and exposure location distance, you may use as a default, the annual dispersion factor ( $0.0045 \mu\text{g}/\text{m}^3$  / pounds/year) and daily dispersion factor ( $4.8 \mu\text{g}/\text{m}^3$  / pounds/day) for a building area of  $\leq 3,000 \text{ ft}^2$ , height of  $\leq 20$  feet, and exposure location distance of 50 meters.

## 3.4 Level 2 Risk Assessment

A Level 2 risk assessment is similar to a Level 1 risk assessment, except that it is less conservative and therefore more accurate. Level 2 assessments are based on air dispersion modeling using AERSCREEN or AERMOD-MAKEMET. AERSCREEN is the easier model to use, and is more appropriate for relatively basic sources with one or a few stacks in flat terrain. AERMOD-MAKEMET can consider the effects of elevated terrain and multiple stacks. Both use conservative screening meteorology. Assuming flat terrain and a single stack, both models will provide the same results. If you plan to conduct modeling, you must develop an air dispersion modeling protocol, and obtain DEQ approval before completing any modeling. DEQ can assist in preparing this simple protocol. A Level 2 risk assessment submitted without an approved modeling protocol may not be accepted.

The model results from AERSCREEN are estimated at the nearest distance from the stack or building to each exposure location, such as a residence. The results from AERMOD-MAKEMET are estimated at specific exposure locations. AERMOD-MAKEMET will provide concentrations at receptors located on a grid, and receptors at specific locations, such as schools or daycare centers, which are important to evaluate in the risk assessment.

Because of the nature of the conservative screening meteorology, Level 2 assessment models only estimate 1-hr concentrations, which must then be converted to daily (24-hour) and annual concentrations. EPA conversion factors should be used to convert the modeled 1-hr concentrations to annual and 24-hour concentrations. This calculation is done automatically for AERSCREEN, but must be done by the user for AERMOD-MAKEMET. These factors to convert 1-hr to 24-hr and annual average concentrations are 0.6 and 0.1, respectively (EPA 2016b).

Once you have calculated exposure concentrations at the various exposure locations, divide the values by the respective RBCs for the toxic air contaminants from OAR 340-245-8010 Table 2, and sum the results as discussed in Section 3.1.2. Compare these calculated total risk values with the RALs in OAR

340-245-8010 Table 1. Example 2 in Appendix A shows an example Level 2 risk assessment.

As with Level 1, calculated risks at Level 2 remain conservative. If further refinement is desired, you can use a Level 3 risk assessment with a more complex model and actual meteorology, which will more accurately characterize toxic air contaminant exposure concentrations.

## 3.5 Level 3 Risk Assessment

The key feature of a Level 3 risk assessment is site-specific air dispersion modeling conducted using a program such as EPA's AERMOD. Because it is important to agree on the modeling receptor grids, appropriate meteorological data, and other elements necessary for effectively running a sophisticated model, DEQ requires that you first submit an air dispersion modeling protocol and risk assessment work plan for DEQ approval prior to conducting the modeling and risk assessment. We recommend that you have at least one meeting with DEQ to agree on scope, and make sure there are common understandings regarding the modeling and risk assessment. DEQ will provide approval of the modeling protocol and work plan in writing. Section 2.6.1 and DEQ's *Recommended Procedures for Air Dispersion Modeling* (DEQ 2022) provide information that will be helpful in preparing the modeling protocol.

Once you have modeled exposure concentrations at the various exposure locations, you can proceed with the calculations as discussed above in Sections 3.1.2 and 3.1.3. In concept, this involves dividing the calculated exposure concentration by the respective RBC in OAR 340-245-8010 Table 2. In practice, DEQ will accept use of one of the REER approaches (C or D). Add the results for all the applicable toxic air contaminants for each of the different types of RBCs (chronic cancer, chronic noncancer, and acute noncancer). By using an air dispersion model such as AERMOD, there is no need to conduct a summation of emissions over TEUs because the aggregate effect of multiple TEUs is already considered. Finally, compare calculated total risk values with the RALs in OAR 340-245-8010 Table 1. Example 3 in Appendix A shows an example Level 3 risk assessment.

## 3.6 Level 4 Risk Assessment

There are two general reasons for conducting Level 4 risk assessments: 1) if you consider it important to incorporate site-specific considerations to more accurately represent risk that may be over-estimated by default assumptions used to develop RBCs, and 2) if DEQ determines that airborne deposition of PBT toxic air contaminants could be important for scenarios not included in the default multipathway adjustment factor (MPAF) assumptions used to develop RBCs. One of DEQ's goals in selecting reasonably protective assumptions for developing RBCs was to minimize the need for Level 4 risk assessments.

DEQ will consider various factors in deciding whether a more detailed multipathway evaluation is required. Generally this evaluation will occur in a Level 4 risk assessment. Some of the factors DEQ will consider are the following:

- Are the emitted chemicals persistent? This includes inorganic chemicals, as well as chemicals with long environmental half-lives, such as PCBs, dioxins, DDT, and other chlorinated pesticides. Toxic air contaminants in Appendix C, Table C-3 with MPAF values are likely the only chemicals that can be quantitatively evaluated for multipathway exposure.
- Are emission rates high enough to result in substantial deposition? This question may be difficult to answer without modeling. It may be informative to make highly protective assumptions such as depositing the entire mass of emitted toxic air contaminants in the area of interest.

- How limited is the food source for humans? Fish in a small pond are unlikely to support substantial fish consumption rates for humans. Similarly, small herds of livestock or dairy cattle are unlikely to support substantial meat or milk consumption rates for humans. A preliminary analysis using conservative air deposition and accumulation assumptions may be sufficient to document that multipathway risks that are unlikely to exceed risk action levels.

A Level 4 risk assessment should have the same elements as Level 3, with some additional considerations. A Level 4 risk assessment should include the following:

- A problem formulation step ending with a conceptual site model identifying TEUs and populations that may be exposed to toxic air contaminant emissions from the source, including residents, nonresident adults, and nonresident children and other sensitive populations;
- A toxicity assessment evaluating the carcinogenicity, noncarcinogenic chronic effects, and noncarcinogenic acute effects of toxic air contaminants to which populations will be exposed, including quantifying noncarcinogenic effects separately for different organ systems, and determining persistence and bioaccumulation potential. Facilities may not consider TRVs other than those listed in OAR 340-247-8010 Table 2 [OAR 340-245-0210(2)(e)];
- An exposure assessment that models or measures toxic air contaminant concentrations at locations of populations that may be exposed to toxic air contaminant emissions from the source. Modifications to default exposure assumptions may be proposed, including but not limited to exposure times, frequencies, and durations, relative bioavailability of toxic air contaminants, and multipathway considerations for PBT toxic air contaminants;
- A risk characterization presenting a quantitative evaluation of potential cumulative health risks associated with exposure to all emissions from the source; and
- A quantitative or qualitative uncertainty evaluation of appropriate elements of the risk assessment.

Elements specific to a Level 4 evaluation include modifications to default exposure assumptions, relative bioavailability of toxic air contaminants, and additional multipathway considerations not addressed by the default adjustment factors. These elements are discussed below.

### **3.6.1 Exposure Assumption Modifications**

The default exposure assumptions for exposure times, frequencies, and durations used in the development of RBCs for residents and workers are typical of those used in risk assessments. However, there may be special circumstances where it is appropriate to modify these assumptions. An example could be a nearby facility that is known to contain workers for only a fraction of the default assumption for exposure time. In this case, you should document the circumstances, and propose modified exposure parameter values for use in the risk assessment.

### **3.6.2 Relative Bioavailability**

The toxicity of a toxic air contaminant can depend on how much of the chemical is actually absorbed by a person, not just on the measured concentration in air (or soil or water). If the form of a toxic air contaminant is less bioavailable to a human than it was in the animal test used as the basis for its TRV, this can be taken into account. For example, under some circumstances, you may want to propose a relative bioavailability test to quantify these differences. There are few standard laboratory tests, and animal tests can be time-consuming and expensive, so relative bioavailability tests are not commonly performed. If you decide to pursue testing, DEQ will request a detailed work plan for approval prior to conducting the evaluation.

Consideration of the particular form of a toxic air contaminant is the main reason for differences in bioavailability and is usually considered when the TRV is established. For example, DEQ has an RBC

for chromium based on the toxicity of hexavalent chromium. The hexavalent form of chromium is substantially more toxic than the other forms, such as trivalent chromium (for which no RBC is available). If you can characterize the specific chemical form of your emissions, you can use the appropriate RBC for that chemical form of the toxic air contaminant at any risk evaluation level. This may make it unnecessary to proceed to a Level 4 evaluation.

### **3.6.3 Multipathway Analysis**

If your source emits PBT toxic air contaminants, it may be important to evaluate air deposition and additional exposure scenarios not included in the development of RBCs that could include contact with soil and water (see Section 2.5). Contact DEQ to discuss how to proceed. Available protocols include EPA's Risk Assessment Guidance for Superfund (1989), Guidelines for Exposure Assessment (1992), and more specifically for toxic air contaminant emissions, California's OEHHA's Air Toxics Hot Spots Program Guidance Manual for Preparation of Health Risk Assessments (2015).

### **3.6.4 Example Level 4 CSM**

As an example of a facility for which a Level 4 risk assessment might be required, consider a facility emitting dioxins near a large lake used as both a drinking water source as well as for recreational fishing. The deposition of dioxins into lake water and subsequent partitioning into sediment will result in ongoing contamination of drinking water that will increase over time. In addition, fish will be contaminated because dioxins in water and sediment will bioaccumulate into fish tissue.

Figure 4 provides an example conceptual site model for the hypothetical facility. In addition to the regular consideration of air exposure to residents, non-residential children, and workers, recreational users of the lake need to be considered. A model such as AERMOD will be required to calculate deposition rates. Given the very slow degradation rates of dioxins, it will be necessary to estimate increases in water and sediment concentrations over time.

## **3.7 Risk Assessment Report**

The risk assessment report should include a summary report of the information provided in the modeling protocol and risk assessment work plan, any information to further support the risk calculations, and the final risk results. The level of detail required in the modeling results varies by the risk assessment level selected. You should provide a written description of the approach, including identification of exposure locations and selection of model input values. Provide sufficient information to allow DEQ to duplicate the results of the modeling and risk assessment during DEQ's review process. The following sections outline the recommendations by risk assessment level. Figure 1 includes a summary of this information.

### **3.7.1. Level 1 and Level 2 AERSCREEN**

Report the following information:

- Provide a map depicting the source location and all relevant exposure locations
- For each toxic air contaminant, provide the RBCs from OAR 340-245-8010 Table 2, dispersion factors from OAR 340-245-8010 Table 3, maximum exposure concentrations, and total excess cancer risk and hazard quotients across all exposure scenarios reported by individual TEU and for the facility as a whole.
- Demonstrate how the total risk across the entire facility was calculated and compared to the risk action levels.

### 3.7.2. Level 2 AERMOD-MAKEMET, Level 3, and Level 4

Report the following information:

- For each toxic air contaminant, provide the RBCs from OAR 340-245-8010 Table 2, location of maximum exposure concentration, maximum exposure concentration, total excess cancer risk and hazard quotients across all exposure scenarios, reported for both TEUs and for the facility as a whole.
- Demonstrate how the total risk across the entire facility was calculated and compared to the risk action levels.
- Provide all modeling input and output files to DEQ. Specifically, DEQ requests the following files:
  - AERMOD input file
  - AERMOD source and receptor files (SOU and ROU)
  - Terrain data files
  - BPIP files
  - Met data (sfc and pfl files)
  - Submit other modeling files needed for running input file
  - Table listing any referenced receptor IDs, geocoordinates (UTM, lat/long), and assigned exposure location.

In addition to the information above, DEQ requests the following information. If a facility does not provide this data, DEQ may create it to better understand the risk near a facility.

- Provide figures showing the concentration/risk plots and gradients around the facility for each exposure scenario.
- For modeling risk using the REER approach, present results in units of risk. Isopleths should represent total risk for each exposure scenario.

DEQ's air dispersion modeling procedures document (DEQ 2022) provides examples of requested report figures. DEQ prefers that contour plots show total facility risk for each exposure scenario. Figure 2 is an example of a figure showing non-cancer acute hazard index for a facility.

### 3.7.3. Risk Assessment Results

Appendix A provides example tables for presenting the results of Level 1, 2, and 3 risk assessments. The summary tables in the risk assessment report should identify locations of maximum risk, and show a direct comparison of calculated risk values with appropriate risk action levels.

By providing all the supporting information identified above, DEQ will be able to review a facility's submitted risk assessment more efficiently. In addition, the documentation will provide a transparent way to show how a facility is meeting the requirements of the Cleaner Air Oregon program.

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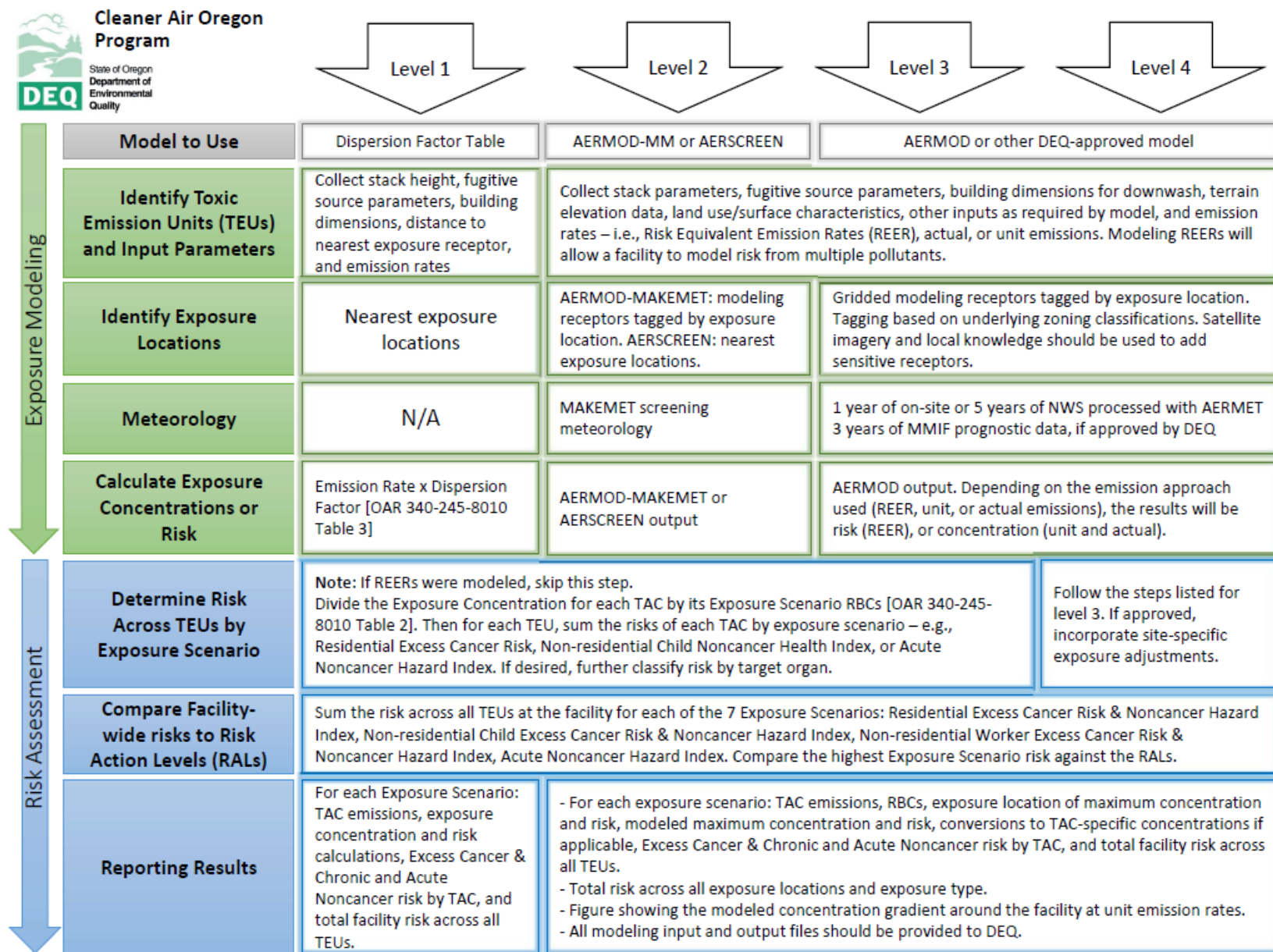
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Figure 1. Exposure Modeling and Risk Assessment Overview

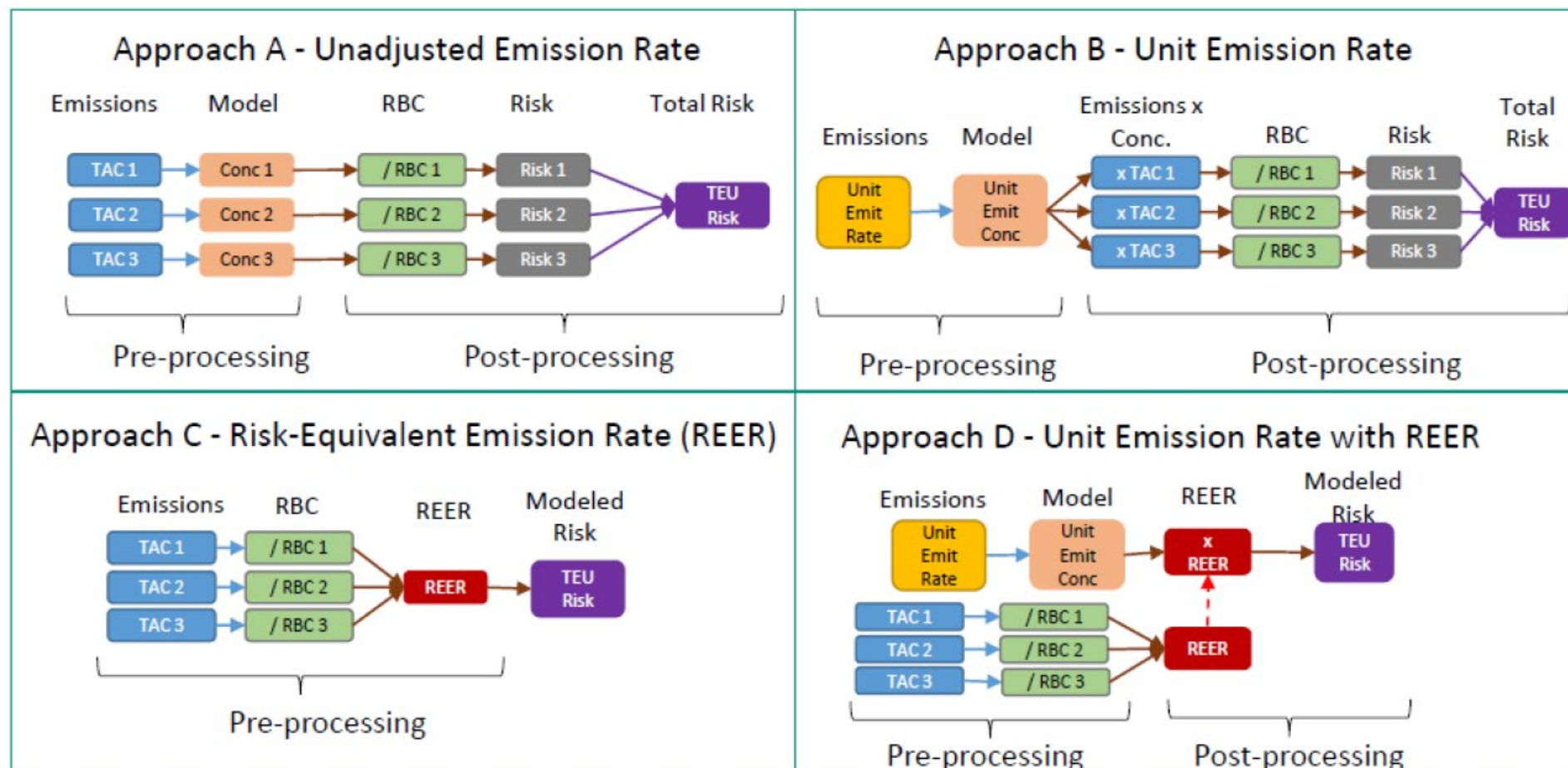




**Figure 2. Example Figure Showing Noncancer Acute Hazard Index**



**Figure 3. Emission Approaches for CAO**



**Definitions:**

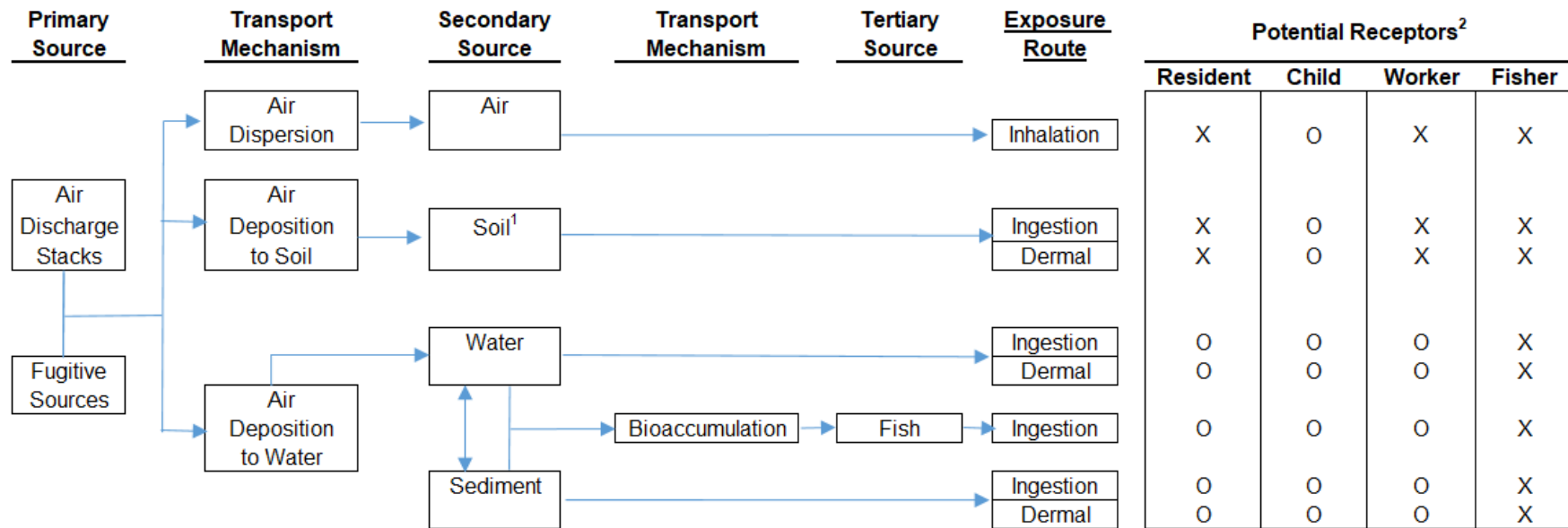
TAC = Toxic Air Contaminant

RBC = Risk-Based Concentration

TEU = Toxics Emission Unit

REER = Risk-Equivalent Emission Rate

**Figure 4. Example Level 4 Conceptual Site Model**



**Notes**

- 1) Multipathway Adjustment Factors incorporated into RBCs include contact with soil and uptake into garden vegetables
- 2) X = complete pathway  
O = incomplete pathway



# APPENDIX A

## Risk Assessment Examples

### A.1 Introduction

The examples in this appendix show how to perform Level 1, Level 2, and Level 3 risk assessments. The first tables show how to conduct the risk assessments according to the explicit procedures presented in rule. Later tables show a more efficient approach that DEQ considers equivalent to the approach presented in rule. See DEQ's air dispersion modeling procedures document (DEQ 2022) for example figures and more information about how to document the air dispersion modeling.

### A.2 Approach A (Unadjusted Emission Rate) and Approach B (Unit Emission Rate)

Example 1 illustrates the different steps in performing a Level 1 risk assessment for a facility with no natural gas, aggregated, or exempt TEUs. Table A-1 shows the stack heights of two emission units at an existing facility, the nearest distances to various exposure locations, and the corresponding dispersion factors obtained from OAR 340-245-8010 Table 3. Table A-2 shows the calculation of exposure concentrations using the dispersion factors and site-specific emission rates for each chemical. Table A-3 shows the comparison of calculated exposure concentrations with Risk Based Concentrations (RBCs) from OAR 340-245-8010 Table 2, and the resulting risk calculations. The facility emits chemicals with noncancer effects designated as both HI3 and HI5, so a Risk Determination Ratio (RDR) needs to be calculated. An example showing RDR calculations is shown in Table A-16. Note that in Table A-16, both an HI and RDR are calculated. This is because some of the Risk Action Levels (RALs) apply only to noncancer HI, and not RDR. Other RALs apply only to RDR. If an HI does not exceed 3 (the lowest RAL using the RDR approach), the RDR does not need to be calculated. Because the risks in Table A-3 (and Table A-16) exceed RALs for an existing facility, additional evaluations were performed using Level 2 and Level 3 risk assessment procedures.

Table A-4 shows how a Level 2 risk assessment would be done using air concentrations obtained from AERSCREEN air dispersion modeling, the comparison with RBCs, and the resulting risk calculations. Because the facility emits chemicals with noncancer effects designated as both HI3 and HI5, and some of the HI values are above 3, an RDR needs to be calculated, similar to the example in Table A-16 for Level 1. Using more site-specific modeling resulted in more accurate, lower exposure concentrations, but the risks are still above RALs. In practice, the air dispersion model is generally run using a unit emission rate of 1 g/s for each emission unit (Approach B). Exposure concentrations are later calculated in a spreadsheet or database by multiplying the model output (as  $\mu\text{g}/\text{m}^3$  per g/s) by the unadjusted emission rate (g/s).

Table A-5 shows how a Level 3 risk assessment would be done using exposure concentrations obtained from more sophisticated AERMOD air dispersion modeling, the comparison with RBCs, and the resulting risk calculations. In this case, even though the facility emits chemicals with noncancer effects designated as both HI3 and HI5, an RDR does not need to be calculated because the HI values are all below 3. In a Level 3 risk assessment there is no need to sum concentrations resulting from two or more emission sources because the air dispersion model calculates exposure concentrations resulting from all emission sources at the same time. Using the more realistic modeling resulted in more accurate, lower exposure concentrations such that risks are not above RALs.

Table A-6 is a summary of calculated risk in the examples at each risk assessment level.

An owner or operator of a source can start at any level of risk assessment, even Level 4. If the owner or operator starts at Level 1 and finds that risk is greater than Risk Action Levels, they can skip directly to a Level 4 risk assessment if they choose to do so.

### **A.2.1 Consideration of Natural Gas and Aggregated TEUs**

For natural gas and aggregated TEUs, calculated risks need to be reported, but not included in the total source risk. Table A-15 shows the summary for an extension of the Level 1 risk assessment in Example 1. Three TEUs (numbers 4, 5, and 6) contribute only a small amount to risk, so are aggregated and evaluated as one aggregate TEU. TEU 7 is a natural gas boiler with no other emissions. Details behind the calculations of risk for the aggregated and natural gas TEUs are not provided here, but should be provided in a risk assessment report. Tables A-1 and A-2 would be revised accordingly to incorporate details about the additional TEUs.

As shown in Table A-15, the risks from the aggregated TEU are less than the aggregate TEU level RALs for existing sources. Therefore, the risks can be excluded from the total source risk.

There are no RALs for the natural gas combustion TEU, but the risk results need to be provided in the risk assessment report. For transparency, this information is shared with the public.

### **A.2.2 Target Organs**

Noncancer hazards can be evaluated by target organs at any level of risk assessment. In the standard approach, this would involve using the calculated exposure concentrations to calculate a hazard index for each relevant organ. Table A-7 shows an example residential chronic noncancer risk calculation using the information for a Level 2 evaluation in Example 2 (without natural gas or aggregated TEUs). Nickel's chronic TRV applies to two target organ systems, immune and respiratory. A hazard quotient is calculated for each organ system, but the hazard quotients are not additive. Therefore, the sum of the hazard quotients for individual organ systems does not equal the initial hazard index of 12 for residential exposure calculated in Table A-4. Because the facility emits chemicals with noncancer effects designated as both HI3 and HI5, and some of the HI values are above 3, an RDR needs to be calculated. Table A-17 shows an example for residential locations.

Table A-8 shows an example residential exposure hazard index evaluation for target organs based on acute effects. Because the facility emits chemicals with noncancer effects designated as both HI3 and HI5, and one of the HI values is above 3, an RDR needs to be calculated, similar to the example in Table A-17.

If natural gas or aggregated TEUs are present at the facility, Tables A-7 and A-8 can be supplemented accordingly, with tables similar to Table A-15.

### **A.3 Approach C – Risk Equivalent Emission Rate**

The example calculations for the standard approach using two emission units and six chemicals may not be extensive depending on the number of receptors on the modeling grid. DEQ recognizes that results of AERMOD modeling using a unit emission rate of 1 g/s from many emission units for many chemicals will require considerable post-modeling processing to determine the maximum annual average risk and maximum daily risk at each exposure location. With a large number of receptors, the post-modeling risk assessment effort may be onerous. As an alternative, DEQ will accept a mathematically equivalent approach that uses AERMOD to model risk instead of chemical concentrations. This approach will require more model runs, but no post-model calculations are required, which will result in considerable savings in time and effort for the risk assessment.

The key to the alternative approach is to model modified emission rates that incorporate risk, by dividing the emission rate for each toxic air contaminant by the RBC appropriate to an exposure location, creating a Risk Equivalent Emission Rate (REER). Because REERs are directly proportional to risk, REERs from different toxic air contaminants can be summed to calculate a total REER for each toxics emission unit. With this approach, the values AERMOD calculates at each modeling receptor is a unitless value corresponding to either cancer risk per million for carcinogens, or hazard index for noncarcinogens. This is because using REER inputs, AERMOD calculations will convert a toxic air contaminant's g/s emission rate to  $\mu\text{g}/\text{m}^3$  at each receptor. Each exposure scenario and toxic endpoint combination (residential cancer, residential chronic noncancer, nonresidential child cancer, nonresidential child chronic noncancer, worker cancer, worker chronic noncancer, and acute noncancer) will need to be run separately. Noncancer risk evaluation can be further resolved by separate target organ system. If the number of target organs is large, and some of the organ systems do not substantially contribute to risk, the low-risk target organs can be conservatively evaluated in a combined category of "other target organs".

Table A-9 uses the same information in Example 2 (without natural gas or aggregated TEUs), and shows how REERs are calculated using RBCs. REERs are calculated separately for each relevant toxic emission unit. After modeling, the result is in risk units, which makes preparing summary Table A-10 a simple task. For chronic RBCs, the model is run as if annual concentrations are being calculated. For acute RBCs, the model is run as if daily concentrations are being calculated. The maximum values for each exposure location are presented for comparison with RALs.

For facilities emitting both HI3 and HI5 chemicals, two sets of REER calculations are needed: one for comparison with total HI RALs, and one for comparison with RDR RALs. This is because some of the RALs apply only to noncancer HI, and not RDR, and other RALs apply only to RDR, and not HI. Two model runs are therefore needed for Approach C.

As with the standard approach, noncancer hazards can be evaluated by target organs at any level of risk assessment. In the alternative approach, this would first involve calculating REERs for each target organ. Table A-11 shows example REER calculations for residential chronic noncancer risk, and Table A-12 shows the calculations for acute risk. Separate REERs would need to be calculated for the other exposure areas (non-residential child and non-residential worker). Running AERMOD using the REERs for each target organ will create an output that corresponds to hazard index. This will make summarizing risk a simple task of taking the highest hazard index for each exposure location. Table A-13 shows an example summary for residential exposure. Separate tables would be needed for non-residential child and non-residential worker exposure locations.

If natural gas or aggregated TEUs are present at the facility, Tables A-11 and A-12 can be modified accordingly, similar to Table A-15.

#### **A.4 Approach D – Unit Emission Rate with Risk Equivalent Emission Rate**

Some facilities may find it advantageous to use a combination of a unit emission rate with the REER approach. With this hybrid approach, air dispersion modeling is performed using a unit emission rate. In a separate calculation performed in a spreadsheet, unadjusted toxic air contaminant emission rates are divided by RBCs to calculate a REER (Tables A-11 and A-12). For facilities emitting both HI3 and HI5 chemicals, two sets of REER calculations are needed: one for comparison with total HI RALs, and one for comparison with RDR RALs. The REER from this pre-processing step can then be multiplied by the calculated air concentration using the unit emission rate to calculate total risk (risk per million for carcinogens, and hazard index for noncarcinogens). Approach D reduces post-processing efforts while maintaining the flexibility of using unit emission rates.





**Table A-1. Example 1 – Toxics Emissions Unit  
Information and Dispersion Factors**

<b>Toxics Emissions Unit (TEU)</b>	<b>Lookup Parameters<sup>[1]</sup> [meters]</b>		<b>Dispersion Factor<sup>[2]</sup> [conc. / emission rate]</b>
<b>TEU-1</b>	Stack height	10	--
	Distance to:		
	Residential	100	0.00075
	Nonresidential child	200	0.00033
	Nonresidential worker	200	0.00033
	Acute (24-hour)	85	2.7 <sup>[3]</sup>
<b>TEU-2</b>	Stack height	20	--
	Distance to:		
	Residential	150	0.00017
	Nonresidential child	250	0.0001
	Nonresidential worker	250	0.0001
	Acute (24-hour)	135	0.635 <sup>[4]</sup>

**Notes:**

[1] - Lookup parameters include stack height and distance to nearest exposure location type.

[2] - Dispersion factors from OAR 340-245-8010 Table 3.

Units for residential, nonresidential child, and nonresidential worker are [ $\mu\text{g}/\text{m}^3$  per lb/yr].

Units for acute are [ $\mu\text{g}/\text{m}^3$  per lb/day].

[3] - Dispersion Factor interpolated between 2.6 and 2.8  $\mu\text{g}/\text{m}^3$  per lb/day

[4] - Dispersion Factor interpolated between 0.62 and 0.65  $\mu\text{g}/\text{m}^3$  per lb/day



**Table A-2. Example 1 – Level 1 Calculation of Air Concentrations**

				Residential	Nonresidential Child	Nonresidential Worker	Acute (24-hour)
Toxics Emissions Unit				Dispersion Factor <sup>[1]</sup>			
				Annual [µg/m <sup>3</sup> per lb/yr]			Acute [µg/m <sup>3</sup> per lb/day]
TEU-1				0.00075	0.00033	0.00033	2.7
TEU-2				0.00017	0.0001	0.0001	0.635
Toxics Emissions Unit	Toxic Air Contaminant	Emission Rate		Calculated Concentration <sup>[2]</sup>			
		Annual [lb/yr]	Acute <sup>[3]</sup> [lb/day]	Average Annual [µg/m <sup>3</sup> ]			Max Acute [µg/m <sup>3</sup> ]
TEU-1	Cadmium	140	0.38	0.11	0.05	0.05	1.03
	Manganese	70	0.25	0.05	0.02	0.02	0.68
	Nickel (insoluble)	220	0.60	0.17	0.07	0.07	1.62
TEU-2	Acetaldehyde	100,000	300	17	10	10	191
	Acetone	80,000	250	14	8	8	159
	Acrolein	10,000	50	2	1	1	32

**Notes:**

[1] - Dispersion factors from OAR 340-245-8010 Table 3. See Table A-1.

[2] - Concentration = Emission Rate \* Dispersion Factor

[3] - Acute (24-hour) emission rate may be annual rate/365 days, or vary if operation is either less than 365 days/year, or a batch operation.

**Legend:**

blue = calculated cell



Table A-3. Example 1 – Summary Risk Table for Level 1 Risk Assessment

Toxics Emissions Unit	CASRN or DEQ ID <sup>[1]</sup>	Toxic Air Contaminant	Noncancer Class	Residential Exposure					Non-Resident Child Exposure					Non-Resident Worker Exposure					Acute Exposure								
				Annual Conc. <sup>3</sup> [µg/m <sup>3</sup> ]	RBC Cancer [µg/m <sup>3</sup> ]	Excess Cancer Risk <sup>[2]</sup>	RBC Noncancer [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[2]</sup>	Annual Conc. <sup>3</sup> [µg/m <sup>3</sup> ]	RBC Cancer [µg/m <sup>3</sup> ]	Excess Cancer Risk <sup>[2]</sup>	RBC Noncancer [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[2]</sup>	Annual Conc. <sup>3</sup> [µg/m <sup>3</sup> ]	RBC Cancer [µg/m <sup>3</sup> ]	Excess Cancer Risk <sup>[2]</sup>	RBC Noncancer [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[2]</sup>	24-Hour Conc. [µg/m <sup>3</sup> ]	RBC Acute [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[4]</sup>						
TEU-1	7440-43-9	Cadmium	HI3	0.11	0.0006	188	0.005	21	0.046	0.014	3.3	0.037	1.2	0.046	0.0067	6.9	0.037	1.2	1.0	0.03	35						
	7439-96-5	Manganese	HI3	0.05	--	--	0.09	0.58	0.023	--	--	0.4	0.058	0.023	--	0.4	0.058	0.7	0.3	2.3							
	365	Nickel (insoluble)	HI3	0.17	0.0038	43	0.014	12	0.073	0.1	0.7	0.062	1.2	0.073	0.046	1.6	0.062	1.2	1.6	0.2	8.1						
TEU-2	75-07-0	Acetaldehyde	HI3	17	0.45	38	140	0.12	10	12	0.83	620	0.02	10	5.5	1.8	620	0.02	191	470	0.41						
	67-64-1	Acetone	HI3	14	--	--	31,000	0.00044	8	--	--	140,000	0.0001	8	--	--	140,000	0.000057	159	62,000	0.0026						
	107-02-8	Acrolein	HI5	1.7	--	--	0.35	4.86	1	--	--	1.5	0.67	1	--	--	1.5	0.67	32	6.9	4.6						
Totals	Total Unrounded Source Risk (TEU-1 and TEU-2)					268.7		38.3			4.9		3.2			10.3		3.2			49.9						
	Total Rounded Source Risk					269		38			5		3			10		3			50						
Risk Determination Ratio Analysis <sup>[6]</sup>				RDR required:		Yes	RDR =		12.1	RDR required:		Yes	RDR =		1.0	RDR required:		Yes	RDR =		1.0	RDR required:		Yes	RDR =		16.0

Risk Action Levels for Existing Sources	Cancer	Noncancer <sup>(6)</sup>		
		HI=3	HI=5	RDR <sup>(7)</sup>
Source Permit Level	0.5	0.5		
Community Engagement	25	1		
TBACT	50	3	5	1.0
Risk Reduction Level	200	6	10	2.0
Immediate Curtailment	500	12	20	4.0

**Notes:**

[1] - CAS No. is shown unless the contaminant listed includes multiple TACs (such as PAHs), in which case a DEQ ID is shown.

[2] - Excess Cancer Risk = Annual conc. (µg/m<sup>3</sup>) / Cancer RBC (µg/m<sup>3</sup>) expressed as risk per million

[3] - Chronic Hazard Quotient = Annual conc. (µg/m<sup>3</sup>) / Noncancer RBC (µg/m<sup>3</sup>) x 1

[4] - Acute Hazard Quotient = 24-hr conc. (µg/m<sup>3</sup>) / Acute RBC (µg/m<sup>3</sup>) x 1

[5] - If HI exceeds 3, a Risk Determination Ratio (RDR) evaluation table (Example Table A-16) must be included for comparison to RALs if both HI3 and HI5 chemicals are emitted from the facility.

[6] - For noncancer risk, TBACT, Risk Reduction, and Immediate Curtailment levels are dependent on the noncancer class of chemicals emitted by the facility:

If all emissions from the facility are of HI5 chemicals, the RALs are: TBACT, HI = 5; Risk Reduction Level, HI = 10; Immediate Curtailment Level, HI = 20.

If all emissions from the facility are of HI3 chemicals, the RALs are: TBACT, HI = 3; Risk Reduction Level, HI = 6; Immediate Curtailment Level, HI = 12.

If emissions from the facility include a mix of HI5 and HI3 chemicals, the RALs are: TBACT, RDR = 1.0; Risk Reduction Level, RDR = 2.0; Immediate Curtailment Level, RDR = 4.0.

[7] - Risk Determination Ratio calculation is not applicable below TBACT level. For comparison against Source Permit Level and Community Engagement RALs, sum the combined HI3 and HI5 risk and round appropriately [OAR 340-245

[8] -0200].

HI = Hazard Index

RDR = Risk Determination Ratio

RAL = Risk Action Level

RBC = Risk Based Concentration

**Legend:**

blue = calculated cell



Table A-4. Example 2 – Summary Risk Table for Level 2 Risk Assessment

Toxics Emissions Unit	CASRN or DEQ ID <sup>[1]</sup>	Toxic Air Contaminant	Noncancer Class	Residential Exposure					Non-Resident Child Exposure					Non-Resident Worker Exposure					Acute Exposure						
				Annual Conc. [µg/m <sup>3</sup> ]	RBC Cancer [µg/m <sup>3</sup> ]	Excess Cancer Risk <sup>[2]</sup>	RBC Noncancer [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup>	Annual Conc. [µg/m <sup>3</sup> ]	RBC Cancer [µg/m <sup>3</sup> ]	Excess Cancer Risk <sup>[2]</sup>	RBC Noncancer [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup>	Annual Conc. [µg/m <sup>3</sup> ]	RBC Cancer [µg/m <sup>3</sup> ]	Excess Cancer Risk <sup>[2]</sup>	RBC Noncancer [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup>	24-Hour Conc. [µg/m <sup>3</sup> ]	Acute RBC [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[4]</sup>				
TEU-1	7440-43-9	Cadmium	HI3	0.0315	0.00056	56	0.005	6.3	0.0139	0.014	1.0	0.037	0.37	0.0139	0.0067	2.1	0.037	0.37	0.2071	0.03	6.9				
	7439-96-5	Manganese	HI3	0.0158	--	--	0.09	0.18	0.0069	--	--	0.4	0.017	0.0069	--	--	0.4	0.017	0.1350	0.3	0.45				
	365	Nickel (insoluble)	HI3	0.0495	0.0038	13	0.014	3.5	0.0218	0.1	0.2	0.062	0.35	0.0218	0.046	0.5	0.062	0.4	0.3255	0.2	1.6				
	Total TEU-1					69		10			1.2		0.74			2.5		0.74			9.0				
TEU -2	75-07-0	Acetaldehyde	HI3	5.10	0.45	11	140	0.036	3.00	12	0.25	620	0.0048	3.00	5.5	0.5	620	0.0048	38.1	470	0.08				
	67-64-1	Acetone	HI3	4.08	--	--	31,000	0.00013	2.40	--	--	140,000	0.000017	2.40	--	--	140,000	0.000017	31.75	62,000	0.0005				
	107-02-8	Acrolein	HI5	0.51	--	--	0.35	1	0.30	--	--	1.5	0.20	0.30	--	--	1.5	0.20	6.35	6.9	0.92				
	Total TEU-2					11		1.5			0.25		0.2			0.55		0.2			1				
Totals	Total Unrounded Source Risk (TEU-1 and TEU-2)					80.6		11.5			1.46		0.95			3.1		0.95			9.98				
	Total Rounded Source Risk					81		12			1		1			3		1			10				
Risk Determination Ratio Analysis <sup>[5]</sup>						RDR Required?	Yes	RDR=	3.6			RDR Required?	No	RDR=	N/A			RDR Required?	No	RDR=	N/A	RDR Required?	Yes	RDR=	3.2

Risk Action Levels for Existing Sources	Cancer	Noncancer <sup>[6]</sup>		
		HI=3	HI=5	RDR <sup>[7]</sup>
Source Permit Level	0.5			
Community Engagement	25			
TBACT	50	3	5	1.0
Risk Reduction Level	200	6	10	2.0
Immediate Curtailment	500	12	20	4.0

**Notes:**

[1] - CAS No. is shown unless the contaminant listed includes multiple TACs (such as PAHs), in which case a DEQ ID is shown.

[2] - Excess Cancer Risk = Annual conc. (µg/m<sup>3</sup>) / Cancer RBC (µg/m<sup>3</sup>) expressed as risk per million

[3] - Chronic Hazard Quotient = Annual conc. (µg/m<sup>3</sup>) / Noncancer RBC (µg/m<sup>3</sup>) x 1

[4] - Acute Hazard Quotient = 24-hr conc. (µg/m<sup>3</sup>) / Acute RBC (µg/m<sup>3</sup>) x 1

[5] - If HI exceeds 3, a Risk Determination Ratio (RDR) evaluation table (Example Table A-16) must be included for comparison to RALs if both HI3 and HI5 chemicals are emitted from the facility.

[6] - For noncancer risk, TBACT, Risk Reduction, and Immediate Curtailment levels are dependent on the noncancer class of chemicals emitted by the facility:

If all emissions from the facility are of HI5 chemicals, the RALs are: TBACT, HI = 5; Risk Reduction Level, HI = 10; Immediate Curtailment Level, HI = 20.

If all emissions from the facility are of HI3 chemicals, the RALs are: TBACT, HI = 3; Risk Reduction Level, HI = 6; Immediate Curtailment Level, HI = 12.

If emissions from the facility include a mix of HI5 and HI3 chemicals, the RALs are: TBACT, RDR = 1.0; Risk Reduction Level, RDR = 2.0; Immediate Curtailment Level, RDR = 4.0.

[7] - Risk Determination Ratio calculation is not applicable below TBACT level. For comparison against Source Permit Level and Community Engagement RALs, sum the combined HI3 and HI5 risk and round appropriately [OAR 340-245-0200].

HI = Hazard Index

RDR = Risk Determination Ratio

RAL = Risk Action Level

RBC = Risk Based Concentration

**Legend:**

blue = calculated cell



Table A-5. Example 3 – Summary Risk Table for Level 3 Risk Assessment

Toxics Emissions Unit	CASRN or DEQ ID <sup>[1]</sup>	Toxic Air Contaminant	Noncancer Class	Residential Exposure					Non-Resident Child Exposure					Non-Resident Worker Exposure					Acute Exposure			
				Annual Conc. [µg/m <sup>3</sup> ]	RBC Cancer [µg/m <sup>3</sup> ]	Excess Cancer Risk <sup>[2]</sup> [µg/m <sup>3</sup> ]	RBC Noncancer [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup> R2105 R2041	Annual Conc. [µg/m <sup>3</sup> ]	RBC Cancer [µg/m <sup>3</sup> ]	Excess Cancer Risk <sup>[2]</sup> [µg/m <sup>3</sup> ]	RBC Noncancer [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup> C3535 C3501	Annual Conc. [µg/m <sup>3</sup> ]	RBC Cancer [µg/m <sup>3</sup> ]	Excess Cancer Risk <sup>[2]</sup> [µg/m <sup>3</sup> ]	RBC Noncancer [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup> W1121 W1007	24-Hour Conc. [µg/m <sup>3</sup> ]	Acute RBC [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[4]</sup> A1090	
TEU-1	Location of Maximum Risk <sup>[5]</sup>																					
	7440-43-9	Cadmium	HI3	0.0019	0.00056	3.4	0.005	0.38	0.00083	0.014	0.059	0.037	0.022	0.00083	0.0067	0.12	0.037	0.022	0.021	0.03	0.70	
	7439-96-5	Manganese	HI3	0.0009	--	--	0.09	0.010	0.00042	--	--	0.4	0.0011	0.00042	--	--	0.4	0.0011	0.014	0.3	0.047	
	365	Nickel (insoluble)	HI3	0.003	0.0038	0.79	0.014	0.21	0.0013	0.1	0.013	0.062	0.021	0.0013	0.046	0.028	0.062	0.021	0.033	0.2	0.17	
	Total Unit 1			4		0.6			0.07		0.04			0.15		0.044			0.91			
TEU-2	Location of Maximum Risk <sup>[5]</sup>																					
	75-07-0	Acetaldehyde	HI3	0.31	0.45	0.69	140	0.0022	0.18	12	0.015	620	0.00029	0.18	5.5	0.033	620	0.00029	3.8	470	0.0081	
	67-64-1	Acetone	HI3	0.24	--	--	31,000	0.0000077	0.14	--	--	140,000	0.0000010	0.14	--	--	140,000	0.0000010	3.2	62000	0.000052	
	107-02-8	Acrolein	HI5	0.031	--	--	0.35	0.089	0.018	--	--	1.5	0.012	0.018	--	--	1.5	0.012	0.6	6.9	0.087	
	Total Unit 2			1		0.1			0.015		0.012			0.03		0.012			0.10			
Totals	Total Unrounded Source Risk (Unit 1 and Unit 2)			4.87		0.695			0.087		0.057			0.18		0.057			1.01			
	Total Rounded Source (Unit 1 and Unit 2)			5		0.7			0.1		0.1			0.2		0.1			1			
Risk Determination Ratio Analysis <sup>[6]</sup>				RDR Required?	No	RDR=	N/A		RDR Required?	No	RDR=	N/A		RDR Required?	No	RDR=	N/A		RDR Required?	No	RDR=	N/A

Risk Action Levels for Existing Sources	Cancer	Noncancer <sup>[7]</sup>		
		HI=3	HI=5	RDR <sup>[8]</sup>
Source Permit Level	0.5	0.5		
Community Engagement	25	1		
TBACT	50	3	5	1.0
Risk Reduction Level	200	6	10	2.0
Immediate Curtailment	500	12	20	4.0

**Notes:**

[1] - CAS No. is shown unless the contaminant listed includes multiple TACs (such as PAHs), in which case a DEQ ID is shown.

[2] - Excess Cancer Risk = Annual conc. (µg/m<sup>3</sup>) / Cancer RBC (µg/m<sup>3</sup>) expressed as risk per million

[3] - Chronic Hazard Quotient = Annual conc. (µg/m<sup>3</sup>) / Noncancer RBC (µg/m<sup>3</sup>) x 1

[4] - Acute Hazard Quotient = 24-hr conc. (µg/m<sup>3</sup>) / Acute RBC (µg/m<sup>3</sup>) x 1

[5] - UTM locations of maximum risk:

R2105 = 10T 421046 5021013

R2041 = 10T 421040 5021010

C3535 = 10T 421046 5020913

C3501 = 10T 421048 5020920

W1121 = 10T 421040 5020902

W1007 = 10T 421038 5020907

A1090 = 10T 421046 5021028

[6] - If HI exceeds 3, a Risk Determination Ratio (RDR) evaluation table (Example Table A-16) must be included for comparison to RALs if both HI3 and HI5 chemicals are emitted from the facility.

[7] - For noncancer risk, TBACT, Risk Reduction, and Immediate Curtailment levels are dependent on the noncancer class of chemicals emitted by the facility:

If all emissions from the facility are of HI5 chemicals, the RALs are: TBACT, HI = 5; Risk Reduction Level, HI = 10; Immediate Curtailment Level, HI = 20.

If all emissions from the facility are of HI3 chemicals, the RALs are: TBACT, HI = 3; Risk Reduction Level, HI = 6; Immediate Curtailment Level, HI = 12.

If emissions from the facility include a mix of HI5 and HI3 chemicals, the RALs are: TBACT, RDR = 1.0; Risk Reduction Level, RDR = 2.0; Immediate Curtailment Level, RDR = 4.0.

[8] - Risk Determination Ratio calculation is not applicable below TBACT level. For comparison against Source Permit Level and Community Engagement RALs, sum the combined HI3 and HI5 risk and round appropriately [OAR 340-245-0200].

HI = Hazard Index

RDR = Risk Determination Ratio

RAL = Risk Action Level

RBC = Risk Based Concentration

**Legend:**

blue = calculated cell



**Table A-6. Summary of Calculated Risk<sup>[1]</sup> at Different Risk Assessment Levels**

Risk Assessment Level	Residential Exposure			Non-Resident Child Exposure			Non-Resident Worker Exposure			Acute Exposure	
	Excess Cancer Risk	Noncancer Risk		Excess Cancer Risk	Noncancer Risk		Excess Cancer Risk	Noncancer Risk		Noncancer Risk	
		Hazard Index <sup>[2]</sup>	Risk Determination Ratio		Hazard Index <sup>[2]</sup>	Risk Determination Ratio		Hazard Index <sup>[2]</sup>	Risk Determination Ratio	Hazard Index <sup>[2]</sup>	Risk Determination Ratio
Level 1	269	38	12.1	5	3	1.0	10	3	1.0	50	16
Level 2	81	12	3.6	1	1	n/a	3	1	n/a	10	3
Level 3	5	0.7	n/a	0.1	0.1	n/a	0.2	0.1	n/a	1	n/a

Risk Action Levels for Existing Sources	Cancer	Noncancer <sup>[3]</sup>		
		HI=3	HI=5	RDR <sup>[4]</sup>
Source Permit Level	0.5	0.5		
Community Engagement	25	1		
TBACT	50	3	5	1.0
Risk Reduction Level	200	6	10	2.0
Immediate Curtailment	500	12	20	4.0

**Notes:**

[1] - Calculated risks are total calculated risks for the example facility, taken from Examples 1, 2, and 3.

[2] - Summation of all Toxic Air Contaminant HIs (HI=3 and HI = 5)

[3] - For noncancer risk, TBACT, Risk Reduction, and Immediate Curtailment levels are dependent on the noncancer class of chemicals emitted by the facility:

If all emissions from the facility are of HI5 chemicals, the RALs are: TBACT, HI = 5; Risk Reduction Level, HI = 10; Immediate Curtailment Level, HI = 20.

If all emissions from the facility are of HI3 chemicals, the RALs are: TBACT, HI = 3; Risk Reduction Level, HI = 6; Immediate Curtailment Level, HI = 12.

If emissions from the facility include a mix of HI5 and HI3 chemicals, the RALs are: TBACT, RDR = 1.0; Risk Reduction Level, RDR = 2.0; Immediate Curtailment Level, RDR = 4.0.

[4] - Risk Determination Ratio calculation is not applicable below TBACT level.

For comparison against Source Permit Level and Community Engagement RALs, sum the combined HI3 and HI5 risk and round appropriately [OAR 340-245-0200].

HI = Hazard Index

RDR = Risk Determination Ratio

**Legend:**

blue = calculated cell



**Table A-7. Example Level 2 Target Organ Chronic Noncancer Risk Assessment for Residential Exposure**

Toxics Emissions Unit	CASRN or DEQ ID <sup>[1]</sup>	Toxic Air Contaminant	Noncancer Class	Chronic Noncancer Residential Exposure								
				Target Organ <sup>[2]</sup> :	Kidney		Nervous System		Immune System		Respiratory System	
				Annual Concentration [µg/m <sup>3</sup> ]	RBC [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup>	RBC [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup>	RBC [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup>	RBC [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup>
TEU-1	7440-43-9	Cadmium	HI3	0.032	0.005	6.4						
	7439-96-5	Manganese	HI3	0.016			0.09	0.18				
	365	Nickel (insoluble)	HI3	0.05					0.014	3.6	0.014	3.6
	Total Unit 1					6.4		0.18		3.6		3.6
TEU-2	75-07-0	Acetaldehyde	HI3	5							140	0.036
	67-64-1	Acetone	HI3	4			31,000	0.00013				
	107-02-8	Acrolein	HI5	0.5							0.35	1.4
	Total Unit 2							0.00013				1.5
Totals	Total Unrounded Source Risk (Unit 1 and Unit 2)					6.4		0.18		3.6		5.0
	Total Source (Unit 1 and Unit 2)					6		0.2		4		5
Risk Determination Ratio Analysis <sup>[4]</sup>					RDR Required?	Yes	RDR Required?	No	RDR Required?	Yes	RDR Required?	Yes
					RDR=	2.1	RDR=	N/A	RDR=	1.2	RDR=	1.5

Risk Action Levels for Existing Sources	Cancer	Noncancer <sup>[5]</sup>		
		HI=3	HI=5	RDR <sup>[6]</sup>
Source Permit Level	0.5	0.5		
Community Engagement	25	1		
TBACT	50	3	5	1.0
Risk Reduction Level	200	6	10	2.0
Immediate Curtailment	500	12	20	4.0

**Notes:**

[1] - CAS No. is shown unless the contaminant listed includes multiple TACs (such as PAHs), in which case a DEQ ID is shown.

[2] - Hazard indices for specific target organs will not necessarily sum to the total hazard index for all organs.

[3] - Chronic Hazard Quotient = Annual conc. (µg/m<sup>3</sup>) / Noncancer RBC (µg/m<sup>3</sup>) x 1

[4] - If HI exceeds 3, a Risk Determination Ratio (RDR) Evaluation table (Example Table A-16) must be included for comparison to RALs if both HI3 and HI5 chemicals are emitted from the facility.

[5] - For noncancer risk, TBACT, Risk Reduction, and Immediate Curtailment levels are dependent on the noncancer class of chemicals emitted by the facility:

If all emissions from the facility are of HI5 chemicals, the RALs are: TBACT, HI = 5; Risk Reduction Level, HI = 10; Immediate Curtailment Level, HI = 20.

If all emissions from the facility are of HI3 chemicals, the RALs are: TBACT, HI = 3; Risk Reduction Level, HI = 6; Immediate Curtailment Level, HI = 12.

If emissions from the facility include a mix of HI5 and HI3 chemicals, the RALs are: TBACT, RDR = 1.0; Risk Reduction Level, RDR = 2.0; Immediate Curtailment Level, RDR = 4.0.

[6] - Risk Determination Ratio calculation is not applicable below TBACT level.

For comparison against Source Permit Level and Community Engagement RALs, sum the combined HI3 and HI5 risk and round appropriately [OAR 245-340-0200].

HI = Hazard Index

RDR = Risk Determination Ratio

RAL = Risk Action Level

RBC = risk-based concentration

**Legend:**

blue = calculated cell



**Table A-8. Example Level 2 Target Organ Acute Noncancer Risk Assessment for Residential Exposure**

Toxics Emissions Unit	CASRN or DEQ ID <sup>[1]</sup>	Toxic Air Contaminant	Noncancer Class	Acute Noncancer Residential Exposure									
				Target Organ <sup>[2]</sup> :	Eyes		Nervous System		Immune System		Respiratory System		
				Daily Concentration [µg/m <sup>3</sup> ]	RBC [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup>	RBC [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup>	RBC [µg/m <sup>3</sup> ]	Hazard Quotient*	RBC [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup>	
TEU-1	7440-43-9	Cadmium	HI3	0.21								0.03	7.0
	7439-96-5	Manganese	HI3	0.14			0.3	0.47					
	365	Nickel (insoluble)	HI3	0.33					0.2	1.65			
	Total Unit 1							0.47		1.65			7.0
TEU-2	75-07-0	Acetaldehyde	HI3	38	470	0.08						470	0.1
	67-64-1	Acetone	HI3	32			62,000	0.00052					
	107-02-8	Acrolein	HI5	6.4								6.9	0.9
	Total Unit 2							0.08		0.00052			1.0
Totals	Total Unrounded Source Risk (Unit 1 and Unit 2)							0.08		0.47		1.65	8.01
	Total Source (Unit 1 and Unit 2)							0.1		0.5		2	8
Risk Determination Ratio Analysis <sup>[4]</sup>					RDR Required?	No	RDR Required?	No	RDR Required?	No	RDR Required?	Yes	
					RDR=	N/A	RDR=	N/A	RDR=	N/A	RDR=	2.5	

Risk Action Levels for Existing Sources	Cancer	Noncancer <sup>[5]</sup>		
		HI=3	HI=5	RDR <sup>[6]</sup>
Source Permit Level	0.5	0.5		
Community Engagement	25	1		
TBACT	50	3	5	1.0
Risk Reduction Level	200	6	10	2.0
Immediate Curtailment	500	12	20	4.0

**Notes:**

[1] - CAS No. is shown unless the contaminant listed includes multiple TACs (such as PAHs), in which case a DEQ ID is shown.

[2] - Hazard indices for specific target organs will not necessarily sum to the total hazard index for all organs.

[3] - Acute Hazard Quotient = 24-hr conc. (µg/m<sup>3</sup>) / Acute RBC (µg/m<sup>3</sup>) x 1

[4] - If HI exceeds 3, a Risk Determination Ratio (RDR) Evaluation table (Example Table A-16) must be included for comparison to RALs if both HI3 and HI5 chemicals are emitted from the facility.

[5] - For noncancer risk, TBACT, Risk Reduction, and Immediate Curtailment levels are dependent on the noncancer class of chemicals emitted by the facility:

If all emissions from the facility are of HI5 chemicals, the RALs are: TBACT, HI = 5; Risk Reduction Level, HI = 10; Immediate Curtailment Level, HI = 20.

If all emissions from the facility are of HI3 chemicals, the RALs are: TBACT, HI = 3; Risk Reduction Level, HI = 6; Immediate Curtailment Level, HI = 12.

If emissions from the facility include a mix of HI5 and HI3 chemicals, the RALs are: TBACT, RDR = 1.0; Risk Reduction Level, RDR = 2.0; Immediate Curtailment Level, RDR = 4.0.

[6] - Risk Determination Ratio calculation is not applicable below TBACT level.

For comparison against Source Permit Level and Community Engagement RALs, sum the combined HI3 and HI5 risk and round appropriately [OAR 245-340-0200].

HI = Hazard Index

RDR = Risk Determination Ratio

RAL = Risk Action Level

RBC = risk-based concentration

**Legend:**

blue = calculated cell





**Table A-9. Example of Level 3 Emissions Calculation for REER Approach**

Toxics Emissions Unit	CASRN or DEQ ID <sup>[1]</sup>	Toxic Air Contaminant	Emission Rates				Residential				Non-Residential Child				Non-Residential Worker				Acute	
			Annual Emission Rate <sup>[2]</sup>		Daily Emission Rate <sup>[3]</sup>		Cancer		Noncancer		Cancer		Noncancer		Cancer		Noncancer			
			(lb/yr)	(g/s)	(lb/day)	(g/s)	RBC Cancer (µg/m³)	Annual REER (g/s per µg/m³)	RBC Noncancer (µg/m³)	Annual REER (g/s per µg/m³)	RBC Cancer (µg/m³)	Annual REER (g/s per µg/m³)	RBC Noncancer (µg/m³)	Annual REER (g/s per µg/m³)	RBC Cancer (µg/m³)	Annual REER (g/s per µg/m³)	RBC Noncancer (µg/m³)	Annual REER (g/s per µg/m³)	Acute RBC (µg/m³)	Daily REER (g/s per µg/m³)
TEU-1	7440-43-9	Cadmium	140	0.0020	0.38	0.0020	0.00056	3.6	0.005	0.40	0.014	0.14	0.037	0.054	0.0067	0.30	0.037	0.054	0.03	0.066
	7439-96-5	Manganese	70	0.0010	0.25	0.0013	—		0.09	0.011	—		0.4	0.0025	—		0.4	0.0025	0.3	0.0044
	365	Nickel (insoluble)	220	0.0032	0.6	0.0031	0.0038	0.83	0.014	0.23	0.10	0.032	0.062	0.051	0.046	0.069	0.062	0.051	0.2	0.016
	Totals	Cancer and HI REER RDR REER						4.4		0.64		0.18		0.11		0.37		0.11		0.087
TEU-2	75-07-0	Acetaldehyde	100,000	1.44	300	1.57	0.45	3.2	140	0.010	12	0.12	620	0.0023	5.5	0.26	620	0.0023	470	0.0034
	67-64-1	Acetone	80,000	1.15	250	1.31	—		31,000	0.000037	—		140,000	0.0000082	—		140,000	0.000008	62000	0.000021
	107-02-8	Acrolein	10,000	0.14	50	0.26	—		0.35	0.41	—		1.5	0.10	—		1.5	0.10	6.9	0.038
	Totals	Cancer and HI REER RDR REER						3.2		0.42		0.12		0.10		0.26		0.10		0.041
Totals	Cancer and HI REER RDR REER							7.6		1.06		0.3		0.21		0.6		0.21		0.13
	RDR REER							0.30		0.30				0.06		0.06		0.06		0.04

**Notes:**

[1] - CAS No. is shown unless the contaminant listed includes multiple TACs (such as PAHs), in which case a DEQ ID is shown.

[2] - Annual Emission Rate (g/s) = Annual Emission Rate (lb/yr) x 453.6 g/lb / (60 sec/min x 60 min/hr x 24 hr/day x 365 day/yr)

[3] - Daily Emission Rate (g/s) = Daily Emission Rate (lb/day) x 453.6 g/lb / (60 sec/min x 60 min/hr x 24 hr/day)

HI = Hazard Index

RBC = Risk Based Concentration

REER = Risk Equivalent Emission Rate

**Legend:**

blue = calculated cell



**Table A-10. Example Summary Risk Table for Level 3 Risk Assessment using REER Approach**

Location of Maximum Risk <sup>[2]</sup>	Residential			Non-Residential Child			Non-Residential Worker			Acute	
	Excess Cancer Risk	Hazard Index	Risk Determination Ratio	Excess Cancer Risk	Hazard Index	Risk Determination Ratio	Excess Cancer Risk	Hazard Index	Risk Determination Ratio	Hazard Index	Risk Determination Ratio
	R2105	R2041		C3535	C3501		W1121	W1007		A1090	
<b>Total Source (TEU-1 and TEU-2)<sup>[3]</sup></b>	<b>5</b>	<b>0.7</b>	<b>n/a</b>	<b>0.1</b>	<b>0.1</b>	<b>n/a</b>	<b>0.2</b>	<b>0.1</b>	<b>n/a</b>	<b>1</b>	<b>n/a</b>

Risk Action Levels for Existing Sources	Cancer	Noncancer <sup>[5]</sup>		
		HI=3	HI=5	RDR <sup>[6]</sup>
Source Permit Level	0.5	0.5		
Community Engagement	25	1		
TBACT	50	3	5	1.0
Risk Reduction Level	200	6	10	2.0
Immediate Curtailment	500	12	20	4.0

**Notes:**

[1] - Values shown are the maximum outputs from AERMOD for each exposure location using modeled risk equivalent emission rates (Table A-9).

[2] - UTM locations of maximum risk:

R2105 = 10T 421046 5021013

R2041 = 10T 421040 5021010

C3535 = 10T 421046 5020913

C3501 = 10T 421048 5020920

W1121 = 10T 421040 5020902

W1007 = 10T 421038 5020907

A1090 = 10T 421046 5021028

[3] - Because the facility emits a combination of HI5 and HI3 chemicals, the TBACT, Risk Reduction, and Immediate Curtailment levels are based on the Risk Determination Ratio, however, the total risk is below the TBACT level.

n/a = non applicable

**Legend:**

blue = calculated cell



**Table A-11. Example of Level 3 Calculation of Emission Rates for REER Approach,  
Target Organ Chronic Noncancer Risk for Residential Exposure**

Toxics Emissions Unit	CASRN or DEQ ID <sup>[1]</sup>	Toxic Air Contaminant	Annual Emission Rate <sup>[2]</sup>		Cancer		Noncancer Target Organ System							
							Kidney		Nervous System		Immune System		Respiratory System	
			[lb/yr]	[g/s]	RBC [µg/m <sup>3</sup> ]	Annual REER <sup>[3]</sup> [g/s per µg/m <sup>3</sup> ]	RBC [µg/m <sup>3</sup> ]	Annual REER <sup>[3]</sup> [g/s per µg/m <sup>3</sup> ]	RBC [µg/m <sup>3</sup> ]	Annual REER <sup>[3]</sup> [g/s per µg/m <sup>3</sup> ]	RBC [µg/m <sup>3</sup> ]	Annual REER [g/s per µg/m <sup>3</sup> ]	RBC [µg/m <sup>3</sup> ]	Annual REER <sup>[3]</sup> [g/s per µg/m <sup>3</sup> ]
TEU-1	7440-43-9	Cadmium	140	0.0020	0.00056	3.6	0.005	0.40						
	7439-96-5	Manganese	70	0.0010					0.09	0.011				
	365	Nickel (insoluble)	220	0.0032	0.0038	0.8					0.014	0.23	0.014	0.23
	TEU 1 Cancer and HI REER					4.4		0.40		0.011		0.23		0.23
	TEU 1 RDR REER							0.13		0.0037		0.075		0.075
TEU-2	75-07-0	Acetaldehyde	100,000	1.438	0.45	3.2							140	0.01
	67-64-1	Acetone	80,000	1.151					31,000	0.000037				
	107-02-8	Acrolein	10,000	0.144									0.35	0.41
	TEU 2 Cancer and HI REER					3.2		n/a		0.000037		n/a		0.421
	TEU 2 RDR REER							n/a		0.000012		n/a		0.086
Totals	Cancer and HI REER					12.1		0.40		0.01		0.23		0.65
	RDR REER							0.13		0.00		0.08		0.16

**Notes:**

[1] - CAS No. is shown unless the contaminant listed includes multiple TACs (such as PAHs), in which case a DEQ ID is shown.

[2] - Annual Emission Rate (g/s) = Annual Emission Rate (lb/yr) x 453.6 g/lb / (60 sec/min x 60 min/hr x 24 hr/day x 365 day/yr)

[3] - REER (g/s per µg/m<sup>3</sup>) = Annual Emission Rate (g/s) / Chronic RBC (µg/m<sup>3</sup>)

RBC = Risk Based Concentration

REER = Risk Equivalent Emission Rate

RDR = Risk Determination Ratio

HI = Hazard Index

n/a = non applicable

**Legend:**

blue = calculated cell



**Table A-12. Example of Level 3 Calculation of Emission Rates for REER Approach,  
Target Organ Acute Noncancer Risk for Residential Exposure**

Toxics Emissions Unit	CASRN or DEQ ID <sup>[1]</sup>	Toxic Air Contaminant	Daily Emission Rate <sup>[2]</sup> [lb/day] [g/s]		Eyes		Nervous System		Immune System		Respiratory System	
					RBC (µg/m <sup>3</sup> )	Acute REER <sup>[3]</sup> [g/s per µg/m <sup>3</sup> ]	RBC (µg/m <sup>3</sup> )	Acute REER <sup>[3]</sup> [g/s per µg/m <sup>3</sup> ]	RBC (µg/m <sup>3</sup> )	Acute REER <sup>[3]</sup> [g/s per µg/m <sup>3</sup> ]	RBC (µg/m <sup>3</sup> )	Acute REER <sup>[3]</sup> [g/s per µg/m <sup>3</sup> ]
TEU-1	7440-43-9	Cadmium	0.38	0.0020							0.03	0.066
	7439-96-5	Manganese	0.25	0.0013			0.3	0.0044				
	365	Nickel (insoluble)	0.6	0.0031					0.2	0.016		
	Totals	HI REER RDR REER				n/a n/a		0.0044 0.0015		0.016 0.0052		0.066 0.022
TEU-2	75-07-0	Acetaldehyde	300	1.575	470	0.0034					470	0.0034
	67-64-1	Acetone	250	1.312			62,000	0.000021				
	107-02-8	Acrolein	50	0.262							6.9	0.038
	Totals	HI REER RDR REER				0.0034 0.0011		0.000021 0.000007		n/a n/a		0.041 0.0087
Totals	HI REER RDR REER					0.0034 0.0011		0.0044 0.0015		0.016 0.0052		0.11 0.031

**Notes:**

[1] - CAS No. is shown unless the contaminant listed includes multiple TACs (such as PAHs), in which case a DEQ ID is shown.

[2] - Daily Emission Rate (g/s) = Daily Emission Rate (lb/dy) x 453.6 g/lb / (60 sec/min x 60 min/hr x 24 hr/day)

[3] - REER (g/s per µg/m<sup>3</sup>) = Daily Emission Rate (g/s) / Acute RBC (µg/m<sup>3</sup>)

RBC = Risk Based Concentration

REER = Risk Equivalent Emission Rate

RDR = Risk Determination Ratio

HI = Hazard Index

n/a = non applicable

**Legend:**

blue = calculated cell



**Table A-13. Example Summary Residential Risk Table for Level 3 Risk Assessment using REER Approach**

Location of Maximum Risk <sup>[1]</sup> :	Chronic Residential Exposure					Acute Residential Exposure			
	Excess Cancer Risk	Kidney Hazard Index	Nervous System Hazard Index	Immune System Hazard Index	Respiratory System Hazard Index	Eyes Hazard Index	Nervous System Hazard Index	Immune System Hazard Index	Respiratory System Hazard Index
	R2105	R2020	R2062	R2016	R2112	A1001	A1092	A1044	A998
Total Source (TEU-1 and TEU-2) Cancer and Hazard Index	5	0.4	0.01	0.2	0.3	0.008	0.05	0.2	0.8
Total Source (TEU-1 and TEU-2) RDR <sup>[2]</sup>	n/a	n/a	n/a	n/a	n/a	n/a	n/a	n/a	n/a

Risk Action Levels for Existing Sources	Cancer	Noncancer <sup>[3]</sup>		
		HI=3	HI=5	RDR <sup>[4]</sup>
Source Permit Level	0.5	0.5		
Community Engagement	25	1		
TBACT	50	3	5	1.0
Risk Reduction Level	200	6	10	2.0
Immediate Curtailment	500	12	20	4.0

**Notes:**

[1] - Location of Maximum Risk:

R2105 = 10T 421046 5021013  
R2020 = 10T 421040 5021010  
R2062 = 10T 421046 5020913  
R2016 = 10T 421048 5020920  
R2112 = 10T 421040 5020902  
A1001 = 10T 421038 5020907  
A1092 = 10T 421046 5021028  
A1044 = 10T 421053 5021024  
A998 = 10T 421035 5021011

[2] - Risk Determination Ratio needed only if Hazard Index exceeds 3.

[3] - For noncancer risk, TBACT, Risk Reduction, and Immediate Curtailment levels are dependent on the noncancer class of chemicals emitted by the facility:

If all emissions from the facility are of HI5 chemicals, the RALs are: TBACT, HI = 5; Risk Reduction Level, HI = 10; Immediate Curtailment Level, HI = 20.

If all emissions from the facility are of HI3 chemicals, the RALs are: TBACT, HI = 3; Risk Reduction Level, HI = 6; Immediate Curtailment Level, HI = 12.

If emissions from the facility include a mix of HI5 and HI3 chemicals, the RALs are: TBACT, RDR = 1.0; Risk Reduction Level, RDR = 2.0; Immediate Curtailment Level, RDR = 4.0.

[4] - Risk Determination Ratio calculation is not applicable below TBACT level.

For comparison against Source Permit Level and Community Engagement RALs, sum the combined HI3 and HI5 risk and round appropriately [OAR 245-340-0200].

HI = Hazard Index

Excess Cancer Risk and HI values shown are the maximum outputs from AERMOD for the residential exposure location using REER approach.

RDR = Risk Determination Ratio

RAL = Risk Action Level

n/a = non applicable



**Table A-14. Example of Level 3 Emissions Calculation for Modeling Unit Emissions and REER Approach**

		Residential		Non-Residential Child		Non-Residential Worker		Acute
		Cancer	Noncancer	Cancer	Noncancer	Cancer	Noncancer	Noncancer
REER Calculations [g/s per µg/m <sup>3</sup> ] <sup>[1]</sup>								
TEU-1	Cancer	4.4		0.18		0.37		
	HI		0.64		0.11		0.11	0.09
	RDR		0.21		0.04		0.04	0.03
TEU-2	Cancer	3.2		0.12		0.26		
	HI		0.42		0.10		0.10	0.04
	RDR		0.09		0.02		0.02	0.01
Modeling Results [µg/m <sup>3</sup> ]								
TEU-1	Location of Max Conc. <sup>[2]</sup>	R2105		C3535		W1121		A1090
	Concentraion [µg/m <sup>3</sup> ]	0.944		0.412		0.412		10.52
TEU-2	Location of Max Conc. <sup>[2]</sup>	R2041		C3501		W1007		A1090
	Concentraion [µg/m <sup>3</sup> ]	0.216		0.125		0.125		2.30
Risk Values								
Totals	Cancer Risk	5		0.1		0.2		
	HI Risk		0.7		0.1		0.1	1
	RDR Risk		0.2		0.02		0.02	0.3

Risk Action Levels for Existing Sources	Cancer	Noncancer <sup>[3]</sup>		
		HI=3	HI=5	RDR <sup>[4]</sup>
Source Permit Level	0.5	0.5		
Community Engagement	25	1		
TBACT	50	3	5	1.0
Risk Reduction Level	200	6	10	2.0
Immediate Curtailment	500	12	20	4.0

**Notes:**

[1] - See Table A-9 for example of how to calculate REER values

[2] - Location of Maximum Risk:

R2105 = 10T 421046 5021013

R2041 = 10T 421040 5021010

C3535 = 10T 421046 5020913

C3501 = 10T 421048 5020920

W1121 = 10T 421040 5020902

W1007 = 10T 421038 5020907

A1090 = 10T 421046 5021028

[3] - For noncancer risk, TBACT, Risk Reduction, and Immediate Curtailment levels are dependent on the noncancer class of chemicals emitted by the facility:

If all emissions from the facility are of HI5 chemicals, the RALs are: TBACT, HI = 5; Risk Reduction Level, HI = 10; Immediate Curtailment Level, HI = 20.

If all emissions from the facility are of HI3 chemicals, the RALs are: TBACT, HI = 3; Risk Reduction Level, HI = 6; Immediate Curtailment Level, HI = 12.

If emissions from the facility include a mix of HI5 and HI3 chemicals, the RALs are: TBACT, RDR = 1.0; Risk Reduction Level, RDR = 2.0;

Immediate Curtailment Level, RDR = 4.0.

[4] - Risk Determination Ratio calculation is not applicable below TBACT level.

For comparison against Source Permit Level and Community Engagement RALs, sum the combined HI3 and HI5 risk and round appropriately [OAR 245-340-0200].

HI = Hazard Index

RDR = Risk Determination Ratio

RAL = Risk Action Level

REER = Risk Equivalent Emission Rate



Table A-15. Example Summary Risk Table for Level 1 Risk Assessment Showing Exempt EUs

Toxics Emissions Unit	Gas No. or DEQ ID <sup>[1]</sup>	Toxic Air Contaminant	Residential Exposure					Non-Resident Child Exposure					Non-Resident Worker Exposure					Acute Exposure				
			Annual Avg Conc.	RBC Cancer	Excess Cancer Risk <sup>[2]</sup>	RBC Noncancer	Hazard Quotient <sup>[3]</sup>	Annual Average Conc.	RBC Cancer	Excess Cancer Risk <sup>[2]</sup>	RBC Noncancer	Hazard Quotient <sup>[3]</sup>	Annual Average Conc.	RBC Cancer	Excess Cancer Risk <sup>[2]</sup>	RBC Noncancer	Hazard Quotient <sup>[3]</sup>	24-Hour Average Conc.	Acute RBC	Hazard Quotient <sup>[4]</sup>		
			( $\mu\text{g}/\text{m}^3$ )	( $\mu\text{g}/\text{m}^3$ )		( $\mu\text{g}/\text{m}^3$ )		( $\mu\text{g}/\text{m}^3$ )	( $\mu\text{g}/\text{m}^3$ )		( $\mu\text{g}/\text{m}^3$ )		( $\mu\text{g}/\text{m}^3$ )	( $\mu\text{g}/\text{m}^3$ )		( $\mu\text{g}/\text{m}^3$ )		( $\mu\text{g}/\text{m}^3$ )				
Aggregated TEUs (TEU-4, TEU-5, TEU-6)	67-63-0	Isopropyl alcohol	50	--	--	200	0.25	40	--	--	880	0.045	40	--	--	880	0.045	100	3200	0.031		
	108-10-1	Methyl isobutyl ketone	75	--	--	3000	0.025	60	--	--	13,000	0.0046	60	--	--	13,000	0.0046	150	--	--		
Total Aggregated TEUs			0.3					0.05					0.05					0.03				
Natural Gas Boiler (TEU-7), >10 and <100 MMBTU	71-43-2	Benzene	2.9E-04	0.13	2.2E-03	3	9.7E-05	1.5E-04	3.3	4.4E-05	13	1.1E-05	4.8E-04	1.5	3.2E-04	13	3.7E-05	9.7E-04	29	3.3E-05		
	50-00-0	Formaldehyde	6.2E-04	0.17	3.6E-03	9	6.8E-05	3.1E-04	4.3	7.2E-05	40	7.7E-06	1.0E-03	2	5.1E-04	40	2.6E-05	2.1E-03	49	4.2E-05		
	401	PAHs (excluding naphthalene)	5.0E-06	0.000043	1.2E-01	--	--	2.5E-06	0.0016	1.6E-03	--	--	8.3E-06	0.003	2.8E-03	--	--	1.7E-05	--	--		
	50-32-8	Benzo[a]pyrene <sup>[5]</sup>	6.0E-08	--	--	0.002	3.0E-05	3.0E-08	--	--	0.0088	3.4E-06	1.0E-07	--	--	0.0088	1.1E-05	2.0E-07	0.002	1.0E-04		
	91-20-3	Naphthalene	1.5E-05	0.029	5.2E-04	3.7	4.1E-06	7.5E-06	0.76	9.9E-06	16	4.7E-07	2.5E-05	0.35	7.1E-05	16	1.6E-06	5.0E-05	200	2.5E-07		
	75-07-0	Acetaldehyde	1.6E-04	0.45	3.4E-04	140	1.1E-06	7.8E-05	12	6.5E-06	620	1.3E-07	2.6E-04	5.5	4.7E-05	620	4.2E-07	5.2E-04	470	1.1E-06		
	107-02-8	Acrolein	1.4E-04	--	--	0.35	3.9E-04	6.8E-05	--	--	1.5	4.5E-05	2.3E-04	--	--	1.5	1.5E-04	4.5E-04	6.9	6.5E-05		
	7664-41-7	Ammonia	9.0E-01	--	--	500	1.8E-03	4.5E-01	--	--	2200	2.0E-04	1.5E+00	--	--	2200	6.8E-04	3.0E+00	1200	2.5E-03		
	7440-38-2	Arsenic and compounds	1.0E-05	0.000024	4.2E-01	0.00017	5.9E-02	5.0E-06	0.0013	3.8E-03	0.0024	2.1E-03	1.7E-05	0.0006	2.7E-02	0.0024	6.9E-03	3.3E-05	0.2	1.7E-04		
	7440-41-7	Beryllium and compounds	6.0E-07	0.00042	1.4E-03	0.007	8.6E-05	3.0E-07	0.011	2.7E-05	0.031	9.7E-06	1.0E-06	0.005	2.0E-04	0.031	3.2E-05	2.0E-06	0.02	1.0E-04		
	7440-43-9	Cadmium and compounds	5.5E-05	0.00056	9.8E-02	0.005	1.1E-02	2.8E-05	0.014	2.0E-03	0.037	7.4E-04	9.2E-05	0.0067	1.4E-02	0.037	2.5E-03	1.8E-04	0.03	6.1E-03		
	18540-29-9	Chromium VI	7.0E-05	0.000031	2.3E+00	0.083	8.4E-04	3.5E-05	0.0005	6.7E-02	0.88	4.0E-05	1.2E-04	0.001	1.2E-01	0.88	1.3E-04	2.3E-04	0.3	7.8E-04		
	7440-48-4	Cobalt and compounds	4.2E-06	--	--	0.1	4.2E-05	2.1E-06	--	--	0.44	4.8E-06	7.0E-06	--	--	0.44	1.6E-05	1.4E-05	--	--		
	7440-50-8	Copper and compounds	4.3E-05	--	--	--	--	2.1E-05	--	--	--	--	7.1E-05	--	--	--	--	1.4E-04	100	1.4E-06		
	100-41-4	Ethylbenzene	3.5E-04	0.4	8.6E-04	260	1.3E-06	1.7E-04	10	1.7E-05	1100	1.6E-07	5.8E-04	4.8	1.2E-04	1100	5.2E-07	1.2E-03	22,000	5.2E-08		
	110-54-3	Hexane	2.3E-04	--	--	700	3.3E-07	1.2E-04	--	--	3100	3.7E-08	3.8E-04	--	--	3100	1.2E-07	7.7E-04	--	--		
	7439-92-1	Lead and compounds	2.5E-05	--	--	0.15	1.7E-04	1.3E-05	--	--	0.66	1.9E-05	4.2E-05	--	--	0.66	6.3E-05	8.3E-05	0.15	5.6E-04		
	7439-96-5	Manganese and compounds	1.9E-05	--	--	0.09	2.1E-04	9.5E-06	--	--	0.4	2.4E-05	3.2E-05	--	--	0.4	7.9E-05	6.3E-05	0.3	2.1E-04		
	7439-97-6	Mercury and compounds	1.3E-05	--	--	0.077	1.7E-04	6.5E-06	--	--	0.63	1.0E-05	2.2E-05	--	--	0.63	3.4E-05	4.3E-05	0.6	7.2E-05		
	365	Nickel and compounds	1.1E-04	0.0038	2.8E-02	0.014	7.5E-03	5.3E-05	0.1	5.3E-04	0.062	8.5E-04	1.8E-04	0.046	3.8E-03	0.062	2.8E-03	3.5E-04	0.2	1.8E-03		
	7782-49-2	Selenium and compounds	1.2E-06	--	--	--	--	6.0E-07	--	--	--	--	2.0E-06	--	--	--	--	4.0E-06	2	2.0E-06		
	108-88-3	Toluene	1.3E-03	--	--	5000	2.7E-07	6.6E-04	--	--	22,000	3.0E-08	2.2E-03	--	--	22,000	1.0E-07	4.4E-03	7500	5.9E-07		
	7440-62-2	Vanadium (fume or dust)	1.2E-04	--	--	0.1	1.2E-03	5.8E-05	--	--	0.44	1.3E-04	1.9E-04	--	--	0.44	4.4E-04	3.8E-04	0.8	4.8E-04		
	1330-20-7	Xylene (mixture)	9.9E-04	--	--	220	4.5E-06	4.9E-04	--	--	970	5.1E-07	1.6E-03	--	--	970	1.7E-06	3.3E-03	8700	3.8E-07		
Total Natural Gas Combustion			3					0.1					0.08					0.01				

## Notes:

[1] - CAS No. is shown unless the contaminant listed includes multiple TACs (such as PAHs), in which case a DEQ ID is shown.

[2] - Excess Cancer Risk = Annual conc. ( $\mu\text{g}/\text{m}^3$ ) / Cancer RBC ( $\mu\text{g}/\text{m}^3$ ) expressed as risk per million[3] - Chronic Hazard Quotient = Annual conc. ( $\mu\text{g}/\text{m}^3$ ) / Noncancer RBC ( $\mu\text{g}/\text{m}^3$ ) x 1[4] - Acute Hazard Quotient = 24-hr conc. ( $\mu\text{g}/\text{m}^3$ ) / Acute RBC ( $\mu\text{g}/\text{m}^3$ ) x 1

[5] - Because benzo[a]pyrene is included in PAHs for cancer effects, only noncancer effects need to be evaluated individually for benzo[a]pyrene.

## Legend:

blue = calculated cell



Table A-16. Example of Level 3 Risk Assessment Risk Determination Ratio Evaluation Table

Toxics Emissions Unit	Cas No. or DEQ ID <sup>[1]</sup>	Toxic Air Contaminant	Noncancer Class	Residential Exposure					Non-Resident Child Exposure					Non-Resident Worker Exposure					Acute Exposure		
				Annual Avg Conc. (µg/m <sup>3</sup> )	RBC Cancer (µg/m <sup>3</sup> )	Excess Cancer Risk <sup>[2]</sup>	RBC Noncancer (µg/m <sup>3</sup> )	Hazard Quotient or RDR <sup>[3]</sup>	Annual Average Conc. (µg/m <sup>3</sup> )	RBC Cancer (µg/m <sup>3</sup> )	Excess Cancer Risk <sup>[2]</sup>	RBC Noncancer (µg/m <sup>3</sup> )	Hazard Quotient or RDR <sup>[3]</sup>	Annual Average Conc. (µg/m <sup>3</sup> )	RBC Cancer (µg/m <sup>3</sup> )	Excess Cancer Risk <sup>[2]</sup>	RBC Noncancer (µg/m <sup>3</sup> )	Hazard Quotient or RDR <sup>[3]</sup>	24-Hour Average Conc. (µg/m <sup>3</sup> )	Acute RBC (µg/m <sup>3</sup> )	Hazard Quotient or RDR <sup>[4]</sup>
TEU-1	7440-43-9	Cadmium	HI3	0.11	0.00056	188	0.005	21	0.046	0.014	3.3	0.037	1.2	0.046	0.0067	6.9	0.037	1.2	1.0	0.03	35
	7439-96-5	Manganese	HI3	0.05	--	--	0.09	0.58	0.023	--	--	0.4	0.058	0.023	--	--	0.4	0.058	0.7	0.3	2.3
	365	Nickel (insoluble)	HI3	0.17	0.0038	43	0.014	12	0.073	0.1	0.7	0.062	1.2	0.073	0.046	1.6	0.062	1.2	1.6	0.2	8.1
	Total Unit 1					231		33.37			4		2.48			8		2.48			44.9
TEU-2	75-07-0	Acetaldehyde	HI3	17	0.45	38	140	0.12	10	12	0.83	620	0.016	10	5.5	1.8	620	0.016	191	470	0.41
	67-64-1	Acetone	HI3	14	--	--	31,000	0.00044	8	--	--	140,000	0.000057	8	--	--	140,000	0.000057	159	62000	0.0026
	107-02-8	Acrolein	HI5	1.7	--	--	0.35	4.86	1	--	--	1.5	0.67	1	--	--	1.5	0.67	32	6.9	4.6
	Total Unit 2					38		4.98			1		0.68			2		0.68			5.01
Totals	Total HI 3 Toxic Air Contaminants							33.49					0.016					0.016			45.32
	Total HI 5 Toxic Air Contaminants							4.86					0.67					0.67			4.60
	Total HI							38.35					3.16					3.16			49.9
	Total RDR (Unit 1 and Unit 2) <sup>[5]</sup>							12.14					0.96					0.96			16.03
Total Rounded Cancer Risk (TEU-1 and TEU-2)						269					5					10					
Total Rounded RDR (TEU-1 and TEU-2)								12.1					1.0					1.0			16.0

Risk Action Levels for Existing Sources	Cancer	Noncancer <sup>[6]</sup>		
		HI=3	HI=5	RDR <sup>[7]</sup>
Source Permit Level	0.5	0.5		
Community Engagement	25	1		
TBACT	50	3	5	1.0
Risk Reduction Level	200	6	10	2.0
Immediate Curtailment	500	12	20	4.0

Notes:

[1] - CAS No. is shown unless the contaminant listed includes multiple TACs (such as PAHs), in which case a DEQ ID is shown.

[2] - Excess Cancer Risk = Annual conc. (µg/m<sup>3</sup>) / Cancer RBC (µg/m<sup>3</sup>) expressed as risk per million

[3] - Chronic Hazard Quotient = Annual conc. (µg/m<sup>3</sup>) / Noncancer RBC (µg/m<sup>3</sup>) x 1

[4] - Acute Hazard Quotient = 24-hr conc. (µg/m<sup>3</sup>) / Acute RBC (µg/m<sup>3</sup>) x 1

[5] - If HI exceeds 3, a Risk Determination Ratio (RDR) Evaluation table (Example Table A-16) must be included for comparison to RALs if both HI3 and HI5 chemicals are emitted from the facility.

[6] - For noncancer risk, TBACT, Risk Reduction, and Immediate Curtailment levels are dependent on the noncancer class of chemicals emitted by the facility:

If all emissions from the facility are of HI5 chemicals, the RALs are: TBACT, HI = 5; Risk Reduction Level, HI = 10; Immediate Curtailment Level, HI = 20.

If all emissions from the facility are of HI3 chemicals, the RALs are: TBACT, HI = 3; Risk Reduction Level, HI = 6; Immediate Curtailment Level, HI = 12.

If emissions from the facility include a mix of HI5 and HI3 chemicals, the RALs are: TBACT, RDR = 1.0; Risk Reduction Level, RDR = 2.0; Immediate Curtailment Level, RDR = 4.0.

[7] - Risk Determination Ratio calculation is not applicable below TBACT level. For comparison against Source Permit Level and Community Engagement RALs, sum the combined HI3 and HI5 risk and round appropriately [OAR 340-245-0200].

HI = Hazard Index

RDR = Risk Determination Ratio

RAL = Risk Action Level

Legend:

blue = calculated cell





**Table A-17. Example Level 2 Risk Determination Ratio Evaluation,  
Target Organ Chronic Noncancer Risk for Residential Exposure**

Toxics Emissions Unit	Cas No. or DEQ ID <sup>[1]</sup>	Toxic Air Contaminant	Noncancer Class	Target Organ <sup>[2]</sup> :	Kidney		Nervous System		Immune System		Respiratory System	
				Annual Concentration [µg/m <sup>3</sup> ]	RBC [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup>	RBC [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup>	RBC [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup>	RBC [µg/m <sup>3</sup> ]	Hazard Quotient or Index <sup>[3]</sup>
TEU-1	7440-43-9	Cadmium	HI3	0.032	0.005	6.4						
	7439-96-5	Manganese	HI3	0.016			0.09	0.18				
	365	Nickel (insoluble)	HI3	0.05					0.014	3.6	0.014	3.6
	<b>Total Unit 1</b>					<b>6.4</b>		<b>0.18</b>		<b>3.6</b>		<b>3.6</b>
TEU-2	75-07-0	Acetaldehyde	HI3	5							140	0.036
	67-64-1	Acetone	HI3	4			31,000	0.00013				
	107-02-8	Acrolein	HI5	0.5							0.35	1.4
	<b>Total Unit 2</b>							<b>0.00013</b>				<b>1.5</b>
Totals	Total HI 3 Toxic Air Contaminants					6.40		0.18		3.57		3.61
	Total HI 5 Toxic Air Contaminants					--		--		--		1.43
	Total HI					6.40		0.18		3.57		5.04
	Total Risk Determination Ratio (Unit 1 and Unit 2)					<b>2.13</b>		<b>0.06</b>		<b>1.19</b>		<b>1.49</b>
	<b>Total Rounded Risk Determination Ratio (Unit 1 and Unit 2)<sup>[4]</sup></b>					<b>2.1</b>		<b>0.1</b>		<b>1.2</b>		<b>1.5</b>

Risk Action Levels for Existing Sources	Cancer	Noncancer <sup>[5]</sup>		
		HI=3	HI=5	RDR <sup>[6]</sup>
Source Permit Level	0.5	0.5		
Community Engagement	25	1		
TBACT	50	3	5	1.0
Risk Reduction Level	200	6	10	2.0
Immediate Curtailment	500	12	20	4.0

**Notes:**

[1] - CAS No. is shown unless the contaminant listed includes multiple TACs (such as PAHs), in which case a DEQ ID is shown.

[2] - Hazard indices for specific target organs will not necessarily sum to the total hazard index for all organs.

[3] - Chronic Hazard Quotient = Annual conc. (µg/m<sup>3</sup>) / Noncancer RBC (µg/m<sup>3</sup>) x 1

[4] - If HI exceeds 3, a Risk Determination Ratio (RDR) Evaluation table (Example Table A-16) must be included for comparison to RALs if both HI3 and HI5 chemicals are emitted from the facility.

[5] - For noncancer risk, TBACT, Risk Reduction, and Immediate Curtailment levels are dependent on the noncancer class of chemicals emitted by the facility:

If all emissions from the facility are of HI5 chemicals, the RALs are: TBACT, HI = 5; Risk Reduction Level, HI = 10; Immediate Curtailment Level, HI = 20.

If all emissions from the facility are of HI3 chemicals, the RALs are: TBACT, HI = 3; Risk Reduction Level, HI = 6; Immediate Curtailment Level, HI = 12.

If emissions from the facility include a mix of HI5 and HI3 chemicals, the RALs are: TBACT, RDR = 1.0; Risk Reduction Level, RDR = 2.0; Immediate Curtailment Level, RDR = 4.0.

[6] - Risk Determination Ratio calculation is not applicable below TBACT level.

For comparison against Source Permit Level and Community Engagement RALs, sum the combined HI3 and HI5 risk and round appropriately [OAR 340-245-0200].

HI = Hazard Index

RDR = Risk Determination Ratio

RAL = Risk Action Level

RBC = risk-based concentration

**Legend:**

blue = calculated cell

# APPENDIX B

## Authoritative Sources of Toxicity Reference Values

### B.1 Chronic Values

DEQ used the following authoritative sources of chronic toxicity reference values (TRVs) to make its recommendations to the EQC, which the EQC adopted:

- DEQ Ambient Benchmark Concentrations (ABCs) adopted by the Environmental Quality Commission (EQC)
- EPA Integrated Risk Information System (IRIS) database ([www.epa.gov/iris](http://www.epa.gov/iris))
- EPA Provisional Peer-Reviewed Toxicity Value (PPRTV) database ([www.hhpptv.ornl.gov](http://www.hhpptv.ornl.gov))
- Agency for Toxic Substances and Disease Registry (ATSDR) Toxicological Profiles ([www.atsdr.cdc.gov](http://www.atsdr.cdc.gov))
- California's Office of Environmental Health Hazard Assessment (OEHHA) ([www.oehha.ca.gov](http://www.oehha.ca.gov))

DEQ and OHA selected the most recently published TRV from among the authoritative sources for each toxic air contaminant. This ensures that chronic TRVs are based on the most recent review of scientific studies. Chronic TRVs were developed separately for noncarcinogenic and carcinogenic effects. For cases where DEQ's ABCs were the most recent values, and DEQ's Air Toxics Science Advisory Committee (ATSAC) decided it was inappropriate to develop an ABC based on carcinogenic effects, we did not obtain a cancer TRV from the other authoritative sources. Similarly, if the ATSAC decided it was inappropriate to develop an ABC based on noncarcinogenic effects, DEQ did not obtain a TRV from the other authoritative sources if the ABC was the most recent value.

Inhalation toxicity information for noncancer effects is typically provided as a threshold value, and given different names by different authoritative bodies. For example, EPA calls the value a Reference Concentration (RfC), the federal Agency for Toxic Substances and Disease Registry calls it a Minimal Risk Level (MRL), and California's Office of Environmental Health Hazard Assessment calls it a Risk Exposure Level (REL). For the purposes of this document, all of these will be given the general name "reference concentration" (RfC). For noncancer, Toxicity Reference Values are equal to the Reference Concentrations.

Equation B.1

$$TRV_{\text{noncancer, chronic}} (\mu\text{g}/\text{m}^3) = RfC_{\text{chronic}} (\mu\text{g}/\text{m}^3)$$

Where:

$TRV_{\text{noncancer, chronic}}$  = toxicity reference value for chronic exposures leading to noncancer health effects

RfC = reference concentration for chronic exposures leading to noncancer health effects

Inhalation toxicity values for carcinogens are typically provided as inhalation unit risk (IUR) values. For ease of use in developing RBCs, IURs were converted to TRV concentrations using a consistent target excess cancer risk level of one in one million.

Equation B.2

$$TRV_{\text{cancer}} (\mu\text{g}/\text{m}^3) = \text{Target Risk } (1 \times 10^{-6}) / \text{IUR } (\mu\text{g}/\text{m}^3)^{-1}$$

Where:

$TRV_{\text{cancer}}$  = Toxicity Reference Value for cancer

IUR = Inhalation Unit Risk

## B.2 Acute Values

The approach used to develop acute TRVs is different from the approach for chronic TRVs because fewer authoritative sources create them, and because the authoritative sources make different assumptions about how long people are exposed. Health risks from inhaling toxic air contaminants are the result not only of how concentrated the contaminants are in the air, but also the amount of time people spend breathing them. In CAO risk assessments, DEQ and OHA assume 24 hours of exposure for acute TRVs. CAO is not intended to be a mechanism to address emergencies where exposures of an hour or less could affect health. There are other mechanisms to address emergencies caused by very high accidental releases. Therefore, when making its recommendations to the EQC, DEQ and OHA selected acute TRVs from among authoritative sources by preference for the authoritative source that used assumptions about exposure times that best matched DEQ and OHA's assumed exposure time of 24 hours. The following authoritative sources are listed in order of preference based on how well their TRVs match DEQ and OHA's assumed 24-hours of exposure.

1. DEQ alone or in consultation with DEQ's ATSAC and/or OHA
2. Agency for Toxic Substances and Disease Registry (ATSDR) Toxicological Profiles, Acute Minimal Risk Levels (MRLs)
3. California's Office of Environmental Health Hazard Assessment (OEHHA) Acute Reference Level (REL)
4. Agency for Toxic Substances and Disease Registry (ATSDR) Toxicological Profiles, Intermediate Minimal Risk Levels (MRLs)
5. If no short-term Reference Concentration was available from sources listed here, no short-term Toxicity Reference Value was recommended or proposed

Acute TRVs are only for non-carcinogenic effects. If no short-term toxicity values were available from the above authoritative sources, no short-term TRVs were established. If the short-term TRV was lower than the chronic TRV, the chronic TRV was used for the short-term TRV, because there is generally more confidence in chronic toxicity values. For example, ATSDR's intermediate MRL for vinylidene chloride is  $79 \mu\text{g}/\text{m}^3$ . The chronic noncancer value from IRIS is  $200 \mu\text{g}/\text{m}^3$ . Given the greater confidence in the chronic value, and because it would be inconsistent to have an acute TRV less than a chronic TRV, the chronic noncancer value was used as the acute TRV for vinylidene chloride.

As with chronic noncancer effects, TRVs for acute effects are equal to the RfCs.

Equation B.3

$$TRV_{\text{noncancer, acute}} (\mu\text{g}/\text{m}^3) = RfC_{\text{acute}} (\mu\text{g}/\text{m}^3)$$

Where:

$TRV_{\text{noncancer, acute}}$  = toxicity reference value for acute exposures leading to noncancer health effects

$RfC_{\text{acute}}$  = reference concentration for acute exposures leading to noncancer health effects

# APPENDIX C

## Development of Adjustment Factors and Calculation of Risk-Based Concentrations

### C.1 Introduction

When making its recommendations to the EQC, DEQ calculated risk-based concentrations (RBCs) for the following receptors for chronic exposure:

- Residential, including single family homes, apartments, and condominiums
- Non-residential children, including schools and daycare facilities
- Non-residential adults, including commercial and industrial facilities

DEQ also considered short-term acute exposure.

DEQ made three adjustments of Toxicity Reference Values, if appropriate, to calculate RBCs. The first adjustment is for a scenario-specific consideration of exposure frequency and duration. Another adjustment is for deposition and bioaccumulation of toxic air contaminants, which involve exposure routes other than inhalation alone; this is a multipathway adjustment. The third adjustment considers early-life exposure to toxic air contaminants that exhibit greater toxicity to infants and children. These adjustments are reflected in the chronic RBCs listed in OAR 340-245-8010 Table 2. The development of each adjustment factor is discussed below. None of the adjustment factors is appropriate or necessary for acute RBCs because of the short period of exposure being considered.

DEQ may recommend that the EQC update the RBC tables in this protocol periodically as toxicity values are revised by the authoritative sources. Revised RBCs will be adopted by rule. In addition, DEQ may also recommend that exposure factors may be revised as new information becomes available.

### C.2 Development of Adjustment Factors

#### C.2.1 Scenario-Specific Exposure Frequency and Duration Adjustments

Residential exposure assumes continual, long-term exposure. Because continual, long-term exposure is the basis of most chronic toxicity values, TRVs are most directly appropriate for residential exposure. In this case, no exposure modifications of TRVs are necessary for calculating RBCs. For other exposure, including shorter term, nonresidential child exposure such as at schools, and worker exposure at commercial or industrial facilities, modifications to TRVs are needed to take into consideration the differences in exposure frequency and duration.

For non-residential exposure, DEQ used factors for more limited exposure to calculate RBCs, as follows. For noncarcinogenic effects for either workers or children in schools or daycare, the value of the adjustment factor for childNRAFnC and workerNRAFnC represents someone who is present 8 hrs/day and 250 days/yr (5 days/week for 50 weeks):

#### Equation C.1

$$\text{childNRAFnC} = \text{workerNRAFnC} = (24 \text{ hrs/day} / 8 \text{ hrs/day}) \times (365 \text{ days/yr} / 250 \text{ days/yr}) = 4.4$$

Where:

childNRAFnC = Nonresident adjustment factor, child noncancer (unitless)

workerNRAFnC = Nonresident adjustment factor, worker noncancer (unitless)

These factors apply to chronic RBCs for noncarcinogenic effects because we assume effects may occur after a year of exposure. For carcinogens, we also include factors for exposure duration because we assume nonresidents are not present at one location for an entire lifetime of 70 years. We assume that non-resident children may be exposed from infancy through elementary school, for a total of 12 years. The standard assumption for worker exposure duration is 25 years. The exposure frequency assumption is 250 days/yr (5 days/week for 50 weeks). The NRAF values for cancer effects are:

#### Equation C.2

$$\text{childNRAFc} = (24 \text{ hrs/day} / 8 \text{ hrs/day}) \times (365 \text{ days/yr} / 250 \text{ days/yr}) \times (70 \text{ yrs} / 12 \text{ yrs}) = 26$$

#### Equation C.3

$$\text{workerNRAFc} = (24 \text{ hrs/day} / 8 \text{ hrs/day}) \times (365 \text{ days/yr} / 250 \text{ days/yr}) \times (70 \text{ yrs} / 25 \text{ yrs}) = 12$$

Where:

childNRAFc = Nonresident adjustment factor, child cancer (unitless)

workerNRAFc = Nonresident adjustment factor, worker cancer (unitless)

### C2.1.1. Life Expectancy

When making its recommendation to the EQC, DEQ and OHA decided that the value of 70 years used above is an appropriate estimate of lifetime despite EPA's determination in the 2011 Exposure Factors Handbook that average life expectancy is now 78 years (EPA 2011). A change in lifetime only matters for evaluating carcinogenic effects for less than lifetime exposure (such as workers) because residential exposure is evaluated for a lifetime regardless of duration, and noncancer effects are evaluated in a manner that does not incorporate life expectancy. Considering a change in life expectancy involves deciding between two inconsistencies. A decision to stay with a 70-year life expectancy used in risk assessments since the 1980s is inconsistent with current knowledge. A decision to change to a 78-year life expectancy would make current risk assessments for workers inconsistent with prior risk assessments, even though actual risks have not changed. EPA recommends continued use of a 70-year lifetime for Superfund risk assessments; DEQ's Cleanup Program follows this recommendation. DEQ determined that it is appropriate for Cleaner Air Oregon risk assessments to use a 70-year lifetime, and the EQC adopted that approach. This decision is slightly more protective than assuming a 78-year lifetime.

### C.2.2 Multipathway Adjustment Factors

DEQ considered developing Multipathway Adjustment Factors (MPAFs) specific to Oregon, but determined that the agency had neither the time nor resources to undertake this effort. After

evaluating Multipathway Adjustment Factors from other agencies, including Minnesota (MPCA 2016) and California's South Coast Air Quality Management District (SCAQMD 2016b), DEQ decided to use the Multipathway Adjustment Factors from SCAQMD because of the extensive modeling performed for the development of the Multipathway Adjustment Factors, and the large list of toxic air contaminants evaluated. DEQ acknowledges that exposure conditions may not be the same in Oregon, but considers the MPAFs appropriately protective.

DEQ only applied MPAFs in development of chronic RBCs, and not in development of acute RBCs. Acute RBCs are equal to acute TRVs in OAR 340-247-8010 Table 2. Acute RBCs only consider risks posed by direct inhalation. Assessment of acute risk need not include multipathway analysis.

### C.2.3 Early-Life Adjustment Factors

Carcinogens that harm a cell's genetic material can have greater toxicity during early-life stages such as infancy and early childhood than in adulthood (EPA 2005a). In these cases, we cannot use the cancer Toxicity Reference Value without modification. For most carcinogenic toxic air contaminants acting by a mutagenic mode of action, we use EPA's general approach to account for early-life exposure using age-dependent adjustment factors, ADAFs. The approach is different for two toxic air contaminants. For trichloroethene (TCE), EPA considers early-life appropriate for liver cancer only, and not kidney cancer or non-Hodgkin's Lymphoma. This makes the development of an early-life adjustment factor for TCE more complicated. For vinyl chloride, EPA determined that it should continue to be evaluated using a specific procedure for evaluating early-life exposure. Because of the many details necessary in evaluating early-life exposure, we provide the development of early-life adjustment factors (ELAFs) separately, in Appendix D.

## C.3 Calculation of RBCs

### C.3.1 Residential RBCs

DEQ applied the multipathway adjustment factor (MPAF) and early-life adjustment factor (ELAF) values shown in Table C-1 to the TRVs in OAR 340-247-8010 Table 2 using the following equations to calculate residential risk-based concentrations (RBCs) in OAR 340-245-8010 Table 2.

Equation C.4

$$residRBCc = \frac{TRVc}{ELAFr \cdot MPAFrc}$$

Equation C.5

$$residRBCnc = \frac{TRVnc}{MPAFrnc}$$

Where:

residRBCc = Residential risk-based concentration for cancer effects (µg/m<sup>3</sup>)

residRBCnc = Residential risk-based concentration for noncancer effects (µg/m<sup>3</sup>)

TRVc = Toxicity reference value for cancer effects (µg/m<sup>3</sup>)

TRVnc = Toxicity reference value for noncancer effects (µg/m<sup>3</sup>)

ELAFr = Early-life adjustment factor, resident (unitless)

MPAFrc = multipathway adjustment factor, resident cancer (unitless)

MPAFrnc = multipathway adjustment factor, resident noncancer (unitless)

If multipathway or early-life considerations are not relevant for a toxic air contaminant, these adjustments are omitted. For most toxic air contaminants, this is the case, and the residential RBC is equal to the TRV.

### C.3.2 Non-Residential RBCs

In addition to considerations of MPAF and ELAF for chronic exposure, exposure frequency and exposure duration are also included for non-residential scenarios where exposure will be less than continual exposure for a lifetime. DEQ used the following equations to calculate non-residential RBCs.

Equation C.6

$$nrchildRBCc = \frac{TRVc \cdot childNRAFc}{ELAFnr \cdot MPAFnrc}$$

Equation C.7

$$nrchildRBCnc = \frac{TRVnc \cdot childNRAFnc}{MPAFnrnc}$$

Equation C.8

$$workerRBCc = \frac{TRVc \cdot workerNRAFc}{MPAFnrc}$$

Equation C.9

$$workerRBCnc = \frac{TRVnc \cdot workerNRAFnc}{MPAFnrnc}$$

Where:

nrchildRBCc = Nonresidential child risk-based concentration for cancer effects (µg/m<sup>3</sup>)

nrchildRBCnc = Nonresidential child risk-based concentration for noncancer effects (µg/m<sup>3</sup>)

workerRBCc = Nonresidential worker risk-based concentration for cancer effects (µg/m<sup>3</sup>)

workerRBCnc = Nonresidential worker risk-based concentration for noncancer effects (µg/m<sup>3</sup>)

TRVc = Toxicity reference value for cancer effects (µg/m<sup>3</sup>)

TRVnc = Toxicity reference value for noncancer effects (µg/m<sup>3</sup>)

ELAFnr = Early-life adjustment factor, non-resident (unitless)

MPAFnrc = Multipathway adjustment factor, nonresident cancer (unitless)

MPAFnrnc = Multipathway adjustment factor, nonresident noncancer (unitless)

childNRAFc = Nonresident adjustment factor, child cancer (26) (unitless)

childNRAFnc = Nonresident adjustment factor, child noncancer (4.4) (unitless)

workerNRAFc = Nonresident adjustment factor, worker cancer (12) (unitless)

workerNRAFnc = Nonresident adjustment factor, worker noncancer (4.4) (unitless)



If multipathway or early-life considerations are not relevant for a toxic air contaminant, these adjustments are omitted.

### **C.3.3 Acute RBCs**

The acute Toxicity Reference Value is used directly as the acute Risk-Based Concentration.

Equation C.10

$$acuteRBC = TRVa$$

Where:

acuteRBC = Acute risk-based concentration ( $\mu\text{g}/\text{m}^3$ )

TRVa = Toxicity reference value for acute effects ( $\mu\text{g}/\text{m}^3$ )

<b>Table C-1</b> <b>Adjustment Factors for Risk-Based Concentrations<sup>a,b</sup></b> <b>Multipathway, Early-Life, and Non-Resident Adjustment Factors</b>						
	Chronic Cancer				Chronic Noncancer	
	Early-Life <sup>d</sup>		Multipathway <sup>c</sup>		Multipathway <sup>c</sup>	
Toxic Air Contaminant	Resident ELAFr	Non-Resident ELAFnr	Resident MPAFrc	Non-Resident MPAFnrc	Resident MPAFrc	Non-Resident MPAFnrc
Acrylamide	1.7	4.2				
Arsenic			9.7	4.5	88	28
Benzidine (and its salts)	1.7	4.2				
Benzo[a]pyrene	1.7	4.2	23	6.6		
<i>Bis</i> -(2-ethylhexyl)phthalate (DEHP)			5.2	1		
Cadmium			1	1	2.0	1.2
Chromium VI <sup>g</sup>	1.7	4.2	1.6	1	2.4	1
Coke Oven Emissions	1.7	4.2				
1,2-Dibromo-3-chloropropane (DBCP)	1.7	4.2				
Dichloromethane (methylene chloride)	1.7	4.2				
Ethylene oxide	1.7	4.2				
Fluorides					5.7	2.9
Hexachlorocyclohexanes (mixture)			5.4	1.3	1	1
<i>alpha</i> -Hexachlorocyclohexane			5.4	1.3	1	1
<i>beta</i> -Hexachlorocyclohexane			5.4	1.3	1	1
<i>gamma</i> -Hexachlorocyclohexane			5.4	1.3	1	1
Hydrogen fluoride					6.1	3.0
Lead			11	5.8		
Mercury					3.9	2.1
4,4'-Methylene dianiline (and its dichloride)			7.2	2.5	1	1
Naphthalene			1	1	1	1

<b>Table C-1</b> <b>Adjustment Factors for Risk-Based Concentrations<sup>a,b</sup></b> <b>Multipathway, Early-Life, and Non-Resident Adjustment Factors</b>						
	Chronic Cancer				Chronic Noncancer	
	Early-Life <sup>d</sup>		Multipathway <sup>c</sup>		Multipathway <sup>c</sup>	
Toxic Air Contaminant	Resident ELAFr	Non-Resident ELAFnr	Resident MPAFrc	Non-Resident MPAFnrc	Resident MPAFrc	Non-Resident MPAFnrc
N-Nitrosodiethylamine	1.7	4.2				
N-Nitrosodimethylamine	1.7	4.2				
Polychlorinated biphenyls (PCBs)			19	13	240	11
Polychlorinated biphenyls (PCBs) TEQ <sup>h</sup>			26 <sup>i</sup>	7.6 <sup>i</sup>	310 <sup>i</sup>	6.7 <sup>i</sup>
Polychlorinated dibenzo- <i>p</i> -dioxins (PCDDs) & Polychlorinated dibenzofurans (PCDFs) TEQ <sup>h</sup>			26 <sup>i</sup>	7.6 <sup>i</sup>	310 <sup>i</sup>	6.7 <sup>i</sup>
Polycyclic aromatic hydrocarbons (PAHs)	1.7	4.2	23	6.6		
Trichloroethene (TCE)	1.2 <sup>e</sup>	1.8 <sup>e</sup>				
Urethane (ethyl carbamate)	1.7	4.2				
Vinyl chloride	2 <sup>f</sup>	27 <sup>f</sup>				

**Notes:**

- a Application of adjustments factors in calculating RBCs:

Resident RBC cancer =  $TRV_c / ELAF_r / MPAF_{rc}$

Resident RBC noncancer =  $TRV_{nc} / MPAF_{nrc}$

Non-resident RBC child cancer =  $TRV_c \times childNRAF_c / ELAF_{nr} / MPAF_{nrc}$

Non-resident RBC child noncancer =  $TRV_{nc} \times childNRAF_{nc} / MPAF_{nrc}$

Worker RBC cancer =  $TRV_c \times workerNRAF_c / MPAF_{nrc}$

Worker RBC noncancer =  $TRV_{nc} \times workerNRAF_{nc} / MPAF_{nrc}$

TRV<sub>c</sub> = Toxicity reference value, cancer

TRV<sub>nc</sub> = Toxicity reference value, noncancer

- b Additional adjustment factors:

childNRAF<sub>nc</sub> = Non-residential adjustment factor, noncancer, child = 4.4

workerNRAF<sub>nc</sub> = Non-residential adjustment factor, noncancer, worker = 4.4

Chronic RBCs are based on continual exposure to residents for 70 years. The adjustment for non-resident exposure is:

$(24 \text{ hours/day} / 8 \text{ hours/day}) \times (365 \text{ days/year} / 250 \text{ days/year}) = 4.4$

childNRAF<sub>c</sub> = Non-residential adjustment factor, child, cancer = 26

For carcinogenic effects to children, the non-residential exposure duration assumption is 12 years (infant through elementary school), resulting in a childNRAF<sub>c</sub> value of:

$$(70 \text{ years} / 12 \text{ years}) \times (365 \text{ days/year} / 250 \text{ days/year}) \times (24 \text{ hours/day} / 8 \text{ hours/day}) = 26$$

workerNRAFc = Non-residential adjustment factor, adult worker, cancer = 12

The adjustment for non-resident worker exposure working for 25 years is:

$$(70 \text{ years} / 25 \text{ years}) \times (365 \text{ days/year} / 250 \text{ days/year}) \times (24 \text{ hours/day} / 8 \text{ hours/day}) = 12$$

- c MPAF = multipathway adjustment factor. Sources of multipathway adjustment factors: South Coast Air Quality Management District, Permit Application Package "M", March 2016, Table 8-1.  
South Coast Air Quality Management District, Facility Prioritization Procedures for AB 2588 Program, Nov. 2016, Table 3.  
Toxic air contaminants for which there are MPAFs are considered persistent, bioaccumulative and toxic substances.
- d ELAF = early-life adjustment factor. ELAFs apply to toxic air contaminants determined by EPA to be carcinogens acting by a mutagenic mode of action. The standard ELAF approach is to use age-dependent adjustment factors (ADAFs) of 10 for infants up to 2 years old, and 3 for children aged 2 to 16, unless EPA determines that a chemical-specific approach is appropriate. For applicable toxic air contaminants, ELAFs are incorporated in the derivation of residential and nonresident child RBCs.
- e Early-life adjustment factor for TCE developed by applying ADAFs to one of three toxic endpoints for TCE.
- f Early-life adjustment factor for vinyl chloride developed by assuming exposure during early-life doubles the lifetime cancer risk without early-life exposure. These ELAF values apply to the IUR of  $4.4 \times 10^{-6} (\mu\text{g}/\text{m}^3)^{-1}$  [TRV =  $0.22 \mu\text{g}/\text{m}^3$ ], not the adult/child IUR of  $8.8 \times 10^{-6} (\mu\text{g}/\text{m}^3)^{-1}$  used to calculate the TRV of  $0.11 \mu\text{g}/\text{m}^3$ .
- g Adjustment factors for chromium VI apply to both chromate and dichromate particulates, and chromic acid aerosol mist.
- h TEQ = toxic equivalency (relative to 2,3,7,8-tetrachlorodibenzo-*p*-dioxin)
- i Multipathway adjustment factors are for PCDDs.

# APPENDIX D

## Derivation of Early-Life Adjustment Factors

### D.1 Introduction

This appendix covers the development of early-life adjustment factors (ELAFs) and the evaluation of early-life exposure for certain compounds. For toxic air contaminants that are carcinogens having a mutagenic mode of action, risk may not be fully assessed without incorporation of early-life exposure. Such toxic air contaminants are shown in Table D-1. We provide a general discussion below, with specific evaluations of TCE and vinyl chloride. In the future, as more information becomes available, early-life exposure may need to be considered for other toxic air contaminants. Early-life exposure is included in the derivation of RBCs for residential and non-residential child exposure scenarios.

### D.2 Background

In March 2005, EPA issued new *Guidelines for Carcinogenic Risk Assessment* (EPA 2005a), updating the 1986 guidelines and 1999 interim final guidelines. Also included was *Supplemental Guidance for Assessing Susceptibility from Early-Life Exposure to Carcinogens* (EPA 2005b). In the *Supplemental Guidance*, EPA concluded that some toxic air contaminants, specifically carcinogens acting by a mutagenic mode of action, have a greater cancer impact if exposure occurs during childhood. DEQ included early-life exposure in the derivation of RBCs for the relevant toxic air contaminants. The general approach is to evaluate cancer risk using different adjusted potency factors for three life stages (0 – 2 years, 2 – 16 years, and adult).

EPA created workgroups to provide additional information on how to implement the *Supplemental Guidance*, and provide consistency. One outcome of the workgroups is an EPA memorandum clarifying which toxic air contaminants should be evaluated for early-life exposure (EPA 2006). The list of mutagenic toxic air contaminants, with updates, is provided in Table D-1. Most toxic air contaminants with early-life exposure considerations are evaluated using the default approach. For TCE, EPA considers early-life appropriate for liver cancer only, and not kidney cancer or non-Hodgkin's Lymphoma. Because of this complication, the approach for TCE is discussed separately. In addition, EPA determined that vinyl chloride should continue to be evaluated using a specific procedure for evaluating early-life exposure, so vinyl chloride is also discussed separately.

### D.3 Default Early-Life Adjustment Factors

Risk assessments for carcinogens acting by a mutagenic mode of action (excluding vinyl chloride discussed below) include a term called an age dependent adjustment factor (ADAF) to account for increased carcinogenic potency during early life stages. For ages up to 2 years, the ADAF is 10, indicating a ten-fold increase in carcinogenic potency during this period. For ages from 2 years to 16 years, the ADAF is 3. For ages 16 years and older, the ADAF is 1. Using ADAFs, the differences in potency are incorporated by a factor separate from the inhalation unit risk factor, so only one cancer IUR is needed. In the CAO program, DEQ uses the corresponding TRV rather than the IUR. Risk assessments for carcinogens that do not act by a mutagenic mode of action should be conducted using the TRV without adjustments for age.

In developing exposure parameters for children, EPA decided that it would be more accurate to divide the 2- to 16-year-old stage into two stages (2 to 6 years, and 6 to 16 years). Both stages have the same ADAF value. For inhalation exposure, it is not necessary to separate these age groups, so they are combined in the equation below.

The incorporation of ADAFs is best included in the calculation of age-adjusted exposure duration for inhalation exposure. These factors are used both in forward risk assessments and calculations of RBCs. Equations for age-adjusted intake factors are presented in DEQ's RBDM guidance (DEQ 2003). For carcinogens acting by a mutagenic mode of action, these equations should be modified as follows:

Equation D.1

$$ED_{adj} = ED_2 ADAF_2 + ED_{16} ADAF_{16} + ED_{adult} ADAF_{adult}$$

Where:

ADAF<sub>2</sub> = Age-dependent Adjustment Factor, child 0 to <2 years old (unitless)  
 ADAF<sub>16</sub> = Age-dependent Adjustment Factor, child 2 to <16 years old (unitless)  
 ADAF<sub>adult</sub> = Age-dependent Adjustment Factor, adult (unitless)  
 ED<sub>2</sub> = Exposure duration, child 0 to <2 years old (yr)  
 ED<sub>16</sub> = Exposure duration, child 2 to <16 years old (yr)  
 ED<sub>adult</sub> = Exposure duration, adult (yr)

This approach is discussed in DEQ's risk assessment guidance for the Cleanup Program (DEQ 2010). Default parameter values are shown in Table D-2.

The early-life adjustment factor for residential exposure is the ratio of early-life exposure duration to general exposure.

Equation D.2

$$ELAF_r = (ED_{adj-r} / ED_r) = [(2 \text{ yr} \times 10) + (14 \text{ yr} \times 3) + (54 \text{ yr} \times 1)] / (70 \text{ yr}) = 116 \text{ yr} / 70 \text{ yr} = 1.66$$

Where:

ELAF<sub>r</sub> = Early-life adjustment factor for residential exposure  
 ED<sub>adj-r</sub> = Exposure duration, adjusted for early-life, residential  
 ED<sub>r</sub> = Exposure duration for residential

For nonresidential child exposure, we assume exposure from infancy through elementary school, for a total exposure duration of 12 years. Other factors, such as exposure frequency (250 days/year) and exposure time (8 hours/day), are already accounted for in the non-residential adjustment factor. The nonresidential ELAF is the ratio of early-life exposure to general exposure for the same duration.

Equation D.3

$$ELAF_{nr} = (ED_{adj} \times EF_{nr}) / ED = [(2 \text{ yr} \times 10) + (10 \text{ yr} \times 3)] / 12 \text{ yr} = 50 \text{ yr} / 12 \text{ yr} = 4.2$$

Where:

ELAF<sub>nr</sub> = Early-life adjustment factor for nonresidential exposure

EDadj-nr = Exposure duration, adjusted for early-life, nonresidential

EDnr = Exposure duration for nonresidential

The default ELAF values are applied to the list of toxic air contaminants with early-life adjustments in Table D-1, except for TCE and vinyl chloride, which are addressed using the approaches described below.

#### D.4 Calculation of ELAFs for TCE

One issue that complicates the derivation of RBCs for TCE concerns the incorporation of early-life exposure. There are three cancer endpoints considered in the development of the carcinogenic slope factor and inhalation unit risk (IUR) factor for TCE: kidney cancer, liver cancer, and non-Hodgkin's lymphoma. EPA determined that TCE was carcinogenic by a mutagenic mode of action for kidney cancer (renal cell carcinoma). Accordingly, age-dependent adjustment factors should be used to evaluate early-life exposure to TCE for this endpoint, presumed to be initiated by a mutagenic mode of action. However, EPA did not determine that there is a mutagenic mode of action for the other two cancer endpoints. The precise method for calculating RBCs for TCE is to use slope factors and IURs for each cancer endpoint, determine an RBC for kidney cancer using ADAFs, determine RBCs for liver cancer and non-Hodgkin's lymphoma without assuming early-life exposure, and combine the individual endpoint RBCs to get a comprehensive RBC using the following equation:

Equation D.4

$$RBC_{TCE-total} = \frac{1}{(1/RBC_{TCE-kidney}) + (1/RBC_{TCE-liver}) + (1/RBC_{TCE-lymphoma})}$$

DEQ used this approach to develop the current RBCs for TCE using default exposure assumptions. To develop site-specific RBCs for TCE, we determined RBCs separately for each toxic endpoint, and then combined the RBCs to derive a total RBC as shown above. The toxicity values for each endpoint are the following:

Toxic Endpoint	TCE Inhalation Unit Risk IUR ( $\mu\text{g}/\text{m}^3$ ) <sup>-1</sup>	TCE Toxicity Reference Value TRV ( $\mu\text{g}/\text{m}^3$ )
Kidney cancer	$1.0 \times 10^{-6}$	1
Liver cancer	$1.0 \times 10^{-6}$	1
Non-Hodgkin's lymphoma	$2.1 \times 10^{-6}$	0.476
Total	$4.1 \times 10^{-6}$	0.244

**Note:**

TRV based on one-in-one-million excess cancer risk.

#### D.5 Calculation of ELAFs for Vinyl Chloride

EPA's Integrated Risk Information System (IRIS) report for vinyl chloride includes two derivations of IUR factors, one based on the linearized multistage (LMS) procedure, and one based on the LED<sub>10</sub> approach (EPA 2000). The LED<sub>10</sub> is the lower 95% limit on a dose that is estimated to cause a 10% response. The results are similar, but the LMS approach is used here because that is what is currently used by the EPA regions. For vinyl chloride, LMS values are

slightly less conservative than IUR factors based on the LED<sub>10</sub> approach.

EPA provided IUR factors separately for lifetime exposure as an adult, and lifetime exposure beginning from birth. The values differ by a factor of 2. The unit risk factors provided in IRIS for inhalation exposure are  $4.4 \times 10^{-6}$  risk per  $\mu\text{g}/\text{m}^3$  for adult exposure, and  $8.8 \times 10^{-6}$  risk per  $\mu\text{g}/\text{m}^3$  for adult/child exposure. The Air Toxics Science Advisory Committee chose to use the adult/child IUR in developing an ambient benchmark concentration for vinyl chloride. Because the ATSAC decision on the ABC was selected as the basis for the vinyl chloride TRV, we used the adult/child IUR for developing RBCs. This simplifies the development of an RBC for residential exposure, but complicates a non-residential child RBC.

An example is presented below for the calculation of the inhalation RBC for vinyl chloride. You can use similar concepts in a forward risk assessment.

For vinyl chloride, EPA concludes that because the effects of early-life exposure are qualitatively and quantitatively different from those of later exposures, it is not appropriate to prorate early-life exposures as if they were received at a proportionately lesser rate over a full lifetime. This feature of vinyl chloride toxicity must be considered in the derivation of RBCs for nonresidential exposure. It is already covered in the derivation of the residential RBC.

Following EPA's example, early-life exposure is estimated assuming a lifetime of exposure using the lower (adult) slope factor. For an exposure scenario involving both early-life and additional exposure, the early-life exposure (which is a single value and is not pro-rated for reduced exposure time) is added to a child's nonresidential exposure (which can be pro-rated).

To show explicitly how early-life and adult exposure are incorporated, the following is the general RBC equation:

Equation D.5

$$RBC_{air} = \frac{AT_c \cdot 365 \text{ days/yr} \cdot TRV}{ED \cdot EF_r}$$

Where:

RBC<sub>air</sub> = Risk based concentration for inhalation of air ( $\mu\text{g}/\text{m}^3$ )  
 AT<sub>c</sub> = Averaging time, carcinogens (70 years)  
 ED = Exposure duration (yr)  
 EF<sub>r</sub> = Exposure frequency, residential (365 days/year)  
 TRV = Toxicity reference value ( $\mu\text{g}/\text{m}^3$ )

Because DEQ followed the ATSAC recommendation to develop a vinyl chloride TRV that includes early-life exposure, for this more detailed calculation we multiplied the early-life TRV of  $0.114 \mu\text{g}/\text{m}^3$  by 2 to get a non-early-life TRV of  $0.228 \mu\text{g}/\text{m}^3$ .

The RBC equation was applied separately for early-life exposure, and exposure other than early-life. Early-life exposure is assumed equivalent to a lifetime of adult exposure (70 years).

Equation D.6



$$RBC_{early-life} = \frac{70 \text{ yr} \cdot 365 \frac{\text{days}}{\text{yr}} \cdot 0.228 \mu\text{g}/\text{m}^3}{70 \text{ yr} \cdot 365 \frac{\text{days}}{\text{yr}}}$$

$$= 0.228 \mu\text{g}/\text{m}^3$$

For the other exposure to a nonresidential child, the RBC is:

Equation D.7

$$RBC_{child} = \frac{70 \text{ yr} \cdot 365 \frac{\text{days}}{\text{yr}} \cdot 24 \frac{\text{hrs}}{\text{day}} \cdot 0.228 \mu\text{g}/\text{m}^3}{12 \text{ yr} \cdot 250 \frac{\text{days}}{\text{yr}} \cdot 8 \frac{\text{hrs}}{\text{day}}}$$

$$= 5.83 \mu\text{g}/\text{m}^3$$

Because the definition of "early-life" is not clearly defined for vinyl chloride, including the full non-residential child exposure duration assumption of 12 years for this calculation may slightly overestimate risk.

The RBC for combined exposure as a child and adult is calculated using the following relationship:

Equation D.8

$$\frac{1}{RBC_{early-life/child}} = \frac{1}{RBC_{early-life}} + \frac{1}{RBC_{child}}$$

Equation D.9

$$RBC_{early-life/child} = \frac{1}{\frac{1}{RBC_{early-life}} + \frac{1}{RBC_{child}}}$$

Equation D.10

$$RBC_{early-life/child} = \frac{1}{\frac{1}{0.228} + \frac{1}{5.83}}$$

$$= 0.22 \mu\text{g}/\text{m}^3$$

This same approach can be used for other scenarios in performing a risk assessment for vinyl chloride.

**Table D-1**  
**Toxic Air Contaminant Determined by EPA to be Carcinogens Having a**  
**Mutagenic Mode of Action**

Toxic air contaminant <sup>a</sup>	Chemical Abstract Service Registration Number
Acrylamide	79-06-1
Benzidine	92-87-5
Coke Oven Emissions	
1,2-Dibromo-3-chloropropane (DBCP)	96-12-8
Ethylene oxide	75-21-8
N-Nitrosodiethylamine	55-18-5
N-Nitrosodimethylamine	62-75-9
Polycyclic aromatic hydrocarbons	
Benz[a]anthracene <sup>b</sup>	56-55-3
Benzo[b]fluoranthene <sup>b</sup>	205-99-2
Benzo[k]fluoranthene <sup>b</sup>	207-08-9
Benzo[a]pyrene	50-32-8
Chrysene <sup>b</sup>	218-01-9
Dibenz[a,h]anthracene	53-70-3
Indeno[1,2,3-cd]pyrene <sup>b</sup>	193-39-5
Trichloroethene <sup>c</sup> (TCE)	79-01-6
1,2,3-Trichloropropane	96-18-4
Urethane (ethyl carbamate)	51-79-6
Vinyl chloride <sup>d</sup>	75-01-4

**Notes:**

a) Source: EPA 2006, and EPA Regional Screening Level table (EPA 2022).

b) Although not explicitly included in EPA's list, EPA states that carcinogenic PAHs with a relative potency factor relating the toxicity to the slope factor for benzo[a]pyrene should also be evaluated for early-life exposure.

c) Of the three cancer endpoints considered in the development of the inhalation unit risk (IUR) factor for TCE (kidney cancer, liver cancer, and non-Hodgkin's lymphoma), EPA determined that TCE was carcinogenic by a mutagenic mode of action for kidney cancer (renal cell carcinoma), but not the other endpoints. Age-dependent adjustment factors should be used to evaluate early-life exposure to TCE for kidney cancer, but not the other endpoints.

d) EPA has a specific method for evaluating early-life exposure to vinyl chloride, as presented in EPA's Integrated Risk Information System ([www.epa.gov/iris](http://www.epa.gov/iris)).

**Table D-2**  
**Default Parameter Values for Early-Life Exposure**

Parameter	<2 Years Old	2 to <6 Years Old	6 to <16 Years Old	Adult
ADAF (unitless) <sup>a</sup>	10	3	3	1
ED (yr) <sup>b</sup> residential <sup>c</sup>	2	4	10	54
nonresidential <sup>d</sup>	2	4	6	0
BW (kg) <sup>b</sup>	15	15	80	80
IRS (mg/d) <sup>b</sup>	200	200	100	100
IRW (L/d) <sup>b</sup>	0.78	0.78	2.5	2.5
AF (mg/cm <sup>2</sup> ·event)	0.2	0.2	0.07	0.07
SA (cm <sup>2</sup> ) <sup>b</sup>	2,690	2,690	6,032	6,032
IRA (m <sup>3</sup> /d) <sup>b</sup>	10	10	20	20

**Notes:**

- a) Age-dependent adjustment factor (ADAF) values taken from EPA 2005b.
- b) Exposure values taken from *Exposure Factors Handbook* (EPA 2011), *Risk Assessment Guidance for Superfund* (EPA 1989), and EPA Regional Screening Levels (EPA 2022).  
 ED = exposure duration      BW = body weight  
 IRS = ingestion rate, soil      IRW = ingestion rate, water  
 AF = adherence factor      SA = skin surface area  
 IRA = inhalation rate, air
- c) The standard residential default exposure duration is 70 years.
- d) The nonresidential default exposure duration is 12 years, infancy through elementary school.

# Appendix E

## Use of the Toxic Equivalency Factor Methodology for Dioxins and Furans, PCBs, and PAHs

### E.1 Introduction

The toxicity equivalency factor (TEF) methodology was developed by the U.S. Environmental Protection Agency (EPA) to evaluate the toxicity and assess the risks of a mixture of structurally-related chemicals with a common mechanism of action. Both EPA and the World Health Organization (WHO) use TEFs to evaluate mixtures of polychlorinated dibenzo-*p*-dioxins and polychlorinated dibenzo-*p*-furans (PCDDs and PCDFs) and mixtures of dioxin-like polychlorinated biphenyls (PCBs). TEF methodology specific to mixtures of polycyclic aromatic hydrocarbons (PAHs) are used by the California EPA, the Washington Department of Ecology, and DEQ's Cleanup Program. Further details for each of these three types of mixtures are presented below.

### E.2 Polychlorinated Dibenzo-*p*-dioxins (PCDDs) and Dibenzofurans (PCDFs)

There are 7 distinct PCDD compounds and 10 distinct PCDF compounds, all of which are referred to as congeners. All 17 dioxin/furan congeners are structurally similar and have the same mechanism of toxicity. Because of their similarities, the combined toxicity of these 17 compounds can be estimated using the sum of their doses, which are scaled for potency relative to one component of the mixture for which adequate dose-response toxicity information is available (EPA 2000); this compound is referred to as an "index" chemical. Of these 17 congeners, the toxicity of 2,3,7,8-tetrachlorodibenzo-*p*-dioxin (commonly referred to as 2,3,7,8-TCDD) has been the most extensively studied, and is used as the index chemical.

Each of the congeners is assigned a TEF which represents the relative potency, or toxicity, of each congener to 2,3,7,8-TCDD. Thus, for 2,3,7,8-TCDD, the TEF is 1.0. The TEFs assigned to each dioxin/furan congener are presented in Table E-1.

To evaluate cumulative risk, the concentrations of each of the 17 congeners is multiplied by its specific TEF. Then those 17 adjusted concentrations are summed to produce a Toxic Equivalency, or TEQ, concentration. The TEQ concentration is then compared to the toxicity value for 2,3,7,8-TCDD to determine whether dioxins and furans are present at levels that will cause unacceptable impacts to human health.

The TEF normalization process described above is based on the use of oral toxicity factors. EPA states that TEFs may be applied to other exposure routes, including inhalation, as an interim estimate (*Recommended Toxicity Equivalence Factors (TEFs) for Human Health Risk Assessments of 2,3,7,8-Tetrachlorodibenzo-*p*-dioxin and Dioxin-Like Compounds*, EPA 2010a).

The Risk Based Concentrations for total dioxins and furans, treated as 2,3,7,8-TCDD TEQ, are  $3 \times 10^{-8}$   $\mu\text{g}/\text{m}^3$  for carcinogenic effects, and  $4 \times 10^{-5}$   $\mu\text{g}/\text{m}^3$  for non-carcinogenic effects. These

protective concentrations were obtained using the 1996 OEHHA Inhalation Risk Unit value of 38 per  $\mu\text{g}/\text{m}^3$  for 2,3,7,8-TCDD, and the OEHHA RfC of  $4 \times 10^{-5} \mu\text{g}/\text{m}^3$ , respectively.

In formula form, the TEQ for PCDDs and PCDFs is calculated as:

Equation E.1

$$TEQ_{PCDDs,PCDFs} = \sum_{i=1}^7 (PCDD_i \cdot TEF_i) + \sum_{i=1}^{10} (PCDF_i \cdot TEF_i)$$

### E.3 Polychlorinated Biphenyls (PCBs)

Polychlorinated biphenyls are comprised of a group of 209 congeners, 12 of which are considered dioxin-like in terms of their structural similarity and mechanism of toxicity. Most people are more familiar with the term “PCB Aroclors”. Aroclors are specific mixtures of portions of the 209 PCB congeners (for example, Aroclor 1260, Aroclor 1254), and were created by Monsanto and used commercially to insulate and cool electrical equipment from the 1930s up to 1977, when Monsanto ceased production. EPA banned the use of Aroclors in 1979. However, because PCBs are extremely persistent and bioaccumulate through the food chain, the residual PCBs related to past Aroclor use still exist today, and have spread globally. As Aroclor mixtures deteriorate, their original mixture of PCB congeners changes over time. Therefore, there is a need to be able to evaluate mixtures of PCB congeners rather than Aroclors, although Aroclors are still evaluated in special cases. The term “Total PCBs” is used in different ways, depending on the situation: 1) for a sum of all 209 PCB congeners; 2) for a sum of Aroclors. Additionally and separately, a sum of normalized concentrations of the 12 dioxin-like PCB congeners can be evaluated using a Toxic Equivalency Factor (TEF) methodology.

#### E.3.1 Total PCBs

The Risk Based Concentration for total polychlorinated biphenyls is  $0.01 \mu\text{g}/\text{m}^3$ , and should be compared to a straight summed concentration of all 209 PCB congeners in a mixture.

#### E.3.2 Dioxin-Like PCB Congeners

The 12 dioxin-like PCB congeners are evaluated by applying a TEF methodology, using the dioxin 2,3,7,8-TCDD as the “index” chemical to which the TEQ for the 12 dioxin-like PCB congeners are compared. The 12 dioxin-like congeners are known to be carcinogenic, and typically are assumed to be of more concern than the remaining 197 PCB congeners. Each of the 12 dioxin-like PCB congeners has an assigned TEF (World Health Organization 2005, EPA 2010a); please refer to Table E-2.

To evaluate the concentration of a PCB mixture which contains the dioxin-like congeners, each dioxin-like PCB congener is multiplied by its assigned TEF, and then the results for all 12 are summed to produce a Toxic Equivalency (TEQ) concentration, which is then compared to the toxicity value for 2,3,7,8-TCDD.

Just as with the evaluation of dioxins and furans, the Risk Based Concentrations for the sum of the 12 dioxin-like PCBs, treated as 2,3,7,8-TCDD TEQ, are  $3 \times 10^{-8} \mu\text{g}/\text{m}^3$  for carcinogenic effects, and  $4 \times 10^{-5} \mu\text{g}/\text{m}^3$  for non-carcinogenic effects. These protective concentrations were

obtained using the 1996 OEHHA Inhalation Risk Unit value of 38 per  $\mu\text{g}/\text{m}^3$  for 2,3,7,8-TCDD, and the OEHHA RfC of  $4 \times 10^{-5} \mu\text{g}/\text{m}^3$ , respectively.

In formula form, the TEQ for dioxin-like PCB congeners is calculated as:

Equation E.2

$$\text{TEQ}_{\text{pcb}} = \sum_{i=1}^{12} (\text{PCB}_i \cdot \text{TEF}_i)$$

## E.4 Polycyclic Aromatic Hydrocarbons

Polycyclic aromatic hydrocarbons (PAHs) are produced whenever fossil fuels or organic matter is combusted. PAHs can also exist as contaminants in uncombusted petroleum products. Several PAHs can increase the risk of developing cancers, and one PAH (benzo[a]pyrene) can also impair normal fetal development. Concentrations of individual PAHs should be normalized to a concentration of the PAH benzo[a]pyrene using TEFs. Once this normalization is completed and the TEF results summed, the resulting TEQ concentration can be compared to the toxicity value for benzo[a]pyrene.

The list of 26 PAHs shown in Table E-3 should be used to generate a concentration for total PAHs. Please note that current laboratory analytical methods are available for only a subset of the PAHs in Table E-3.

Because benzo[a]pyrene has both cancer and non-cancer effects, the concentration of benzo[a]pyrene as an individual PAH should also be compared separately against the non-cancer RfC for benzo[a]pyrene, which is an RfC value of  $0.002 \mu\text{g}/\text{m}^3$  (EPA 2017). Because the RfC is based on developmental effects, this value for benzo[a]pyrene should be compared to 24-hour-based concentrations as well as annual averaged concentrations of this PAH.

Naphthalene is both a representative volatile PAH and was the single most emitted PAH in Oregon circa 2005. Thus, at that time, naphthalene was evaluated separately from the other PAHs. Naphthalene is still evaluated separately, and is not included in the summed total carcinogenic PAHs.

## E.5 References

Ecology (Washington Department of Ecology). 2007. Evaluating the toxicity and assessing the carcinogenic risk of environmental mixtures using toxic equivalency factors.

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[https://cfpub.epa.gov/ncea/iris2/chemicalLanding.cfm?substance\\_nmbr=136](https://cfpub.epa.gov/ncea/iris2/chemicalLanding.cfm?substance_nmbr=136)

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EPA. 2009. Technical Assistance Document for the National Air Toxics Trends Stations Program, Revision 2. Prepared for U.S. EPA by Eastern Research Group, Inc., Morrisville, NC. April 1.

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Minnesota Department of Health (MDOH). 2014. Guidance for evaluating the cancer potency of polycyclic aromatic hydrocarbon (PAH) mixtures in environmental samples. February 8.

van den Berg, M; Birnbaum, L; Bosveld, AT; et al. 1998. Toxic equivalency factors (TEFs) for PCBs, PCDDs, PCDFs for humans and wildlife. Environ Health Perspect 106(12): 775-792.

van den Berg, M; Birnbaum, LS; Denison, M: et al. 2006. The 2005 World Health Organization re-evaluation of human and mammalian toxic equivalency factors for dioxins and dioxin-like compounds. Toxicol Sci 93(2): 223-241.

**Table E-1: Toxicity Equivalency Factors for Dioxin/Furan Congeners**

<b>Congener</b>	<b>Toxicity Equivalency Factor<sup>1</sup></b>
<b><i>PCDDs</i></b>	
2,3,7,8-TetraCDD	1
1,2,3,7,8-PentaCDD	1
1,2,3,4,7,8-HexaCDD	0.1
1,2,3,6,7,8-HexaCDD	0.1
1,2,3,7,8,9-HexaCDD	0.1
1,2,3,4,6,7,8-HeptaCDD	0.01
1,2,3,4,6,7,8,9-OctaCDD	0.0003
<b><i>PCDFs</i></b>	
2,3,7,8-TetraCDF	0.1
1,2,3,7,8-PentaCDF	0.03
2,3,4,7,8-PentaCDF	0.3
1,2,3,4,7,8-HexaCDF	0.1
1,2,3,6,7,8-HexaCDF	0.1
2,3,4,6,7,8-HexaCDF	0.1
1,2,3,7,8,9-HexaCDF	0.1
1,2,3,4,6,7,8-HeptaCDF	0.01
1,2,3,4,7,8,9-HeptaCDF	0.01
1,2,3,4,6,7,8,9-OctaCDF	0.0003

**Note**

- 1) From Van den Berg et al. (2006); adopted for use by the World Health Organization and by USEPA (2010).



**Table E-2: Dioxin-Like PCB Congeners and Related TEFs**

<b>Congener</b>	<b>TEF<sup>1</sup></b>
3,3',4,4'-Tetrachlorinated biphenyl (PCB 77)	0.0001
3,4,4',5-TetraCB (PCB 81)	0.0003
3,3',4,4',5-PentaCB (PCB 126)	0.1
3,3',4,4',5,5'-HexaCB (PCB 169)	0.03
2,3,3',4,4'-PentaCB (PCB 105)	0.00003
2,3,4,4',5-PentaCB (PCB 114)	0.00003
2,3',4,4',5-PentaCB (PCB 118)	0.00003
2',3,4,4',5-PentaCB (PCB 123)	0.00003
2,3,3',4,4',5-HexaCB (PCB 156)	0.00003
2,3,3',4,4',5'-HexaCB (PCB 157)	0.00003
2,3',4,4',5,5'-HexaCB (PCB 167)	0.00003
2,3,3',4,4',5,5'-HeptaCB (PCB 189)	0.00003

**Note**

- 1) From Van den Berg et al. (2006); adopted for use by the World Health Organization and by USEPA (2010).

**Table E-3. Recommended Revised List of PAHs and Related TEFs  
(Reduction from 2005 list of 32 PAHs to proposed 26)**

#	PAH	EPA Required (1) <sup>a</sup>	EPA Requested (14) <sup>b</sup>	From MN list (11) <sup>c</sup>	TEF <sup>c,d</sup>
1	5-Methylchrysene			●	1 <sup>d</sup>
2	6-Nitrochrysene			●	10 <sup>d</sup>
3	Acenaphthene		●		NA
4	Acenaphthylene		●		NA
5	Anthanthrene			●	0.4
6	Anthracene		●		0
7	Benz[a]anthracene		●		0.2
8	Benzo[a]pyrene	●			1
9	Benzo[b]fluoranthene		●		0.8
10	Benzo[c]fluorene			●	20
11	Benzo[e]pyrene		●		NA
12	Benzo[g,h,i]perylene			●	0.009
13	Benzo[j]fluoranthene			●	0.3
14	Benzo[k]fluoranthene		●		0.03
15	Chrysene		●		0.1
16	Cyclopenta[c,d]pyrene			●	0.4
17	Dibenz[a,h]anthracene		●		10
18	Dibenzo[a,e]pyrene			●	0.4
19	Dibenzo[a,h]pyrene			●	0.9
20	Dibenzo[a,i]pyrene			●	0.6
21	Dibenzo[a,l]pyrene			●	30
22	Fluoranthene		●		0.08
23	Fluorene		●		NA
24	Indeno[1,2,3-c,d]pyrene		●		0.07
25	Phenanthrene		●		0
26	Pyrene		●		0

**Notes:**

<sup>a</sup> Naphthalene is also required, but already has its own risk-based concentration.

<sup>b</sup> Per EPA National Air Toxics Trend Sites (NATTS) Technical Assistance Document (TAD) 2009, Revision 2, Table 1.1-1. Note that the most-current version of NATTS, published in 2016, requests the same list of PAHs as those presented in the 2009 NATTS.

<sup>c</sup> PAHs on Minnesota Department of Health 2014 list of 19 priority cPAHs that are not already required or requested by EPA.

<sup>d</sup> Values were obtained from an External Review Draft version of EPA's 2010 *Development of a relative potency factor (RPF) approach for polycyclic aromatic hydrocarbon (PAH) mixtures*. Although this document is not supposed to be cited or quoted, the Air Toxics Science Advisory Committee considers this information to be the best and most current science available on this topic. A portion of the TEFs represent the average range of Potency Equivalency Factors provided in this document.

NA – not listed in either EPA 2010b or by MnDOH, but is a NATTS-requested PAH.

# APPENDIX F

## Compilation of Target Organs for Toxic Air Contaminants

You can refine the evaluation of noncancer hazard index for your facility by summing hazard quotients of toxic air contaminants with effects on the same organ system, rather than summing effects on all organ systems. DEQ and OHA determined that the health effect, or health effects, used to derive the TRV should be considered when determining whether to include a toxic air contaminant in an organ-specific hazard index evaluation. Chemicals may cause effects on other organs at concentrations higher than the TRV, but there is greater uncertainty associated with quantifying these effects, and it is not necessary to include them in the analysis. Some TRVs are set based on impacts to more than one target organ. In these cases, assess risk to all target organs that are the basis for the TRV. Table F-1 (chronic effects) and Table F-2 (acute effects) show applicable organ systems for toxic air contaminants that will be acceptable to DEQ for organ-specific evaluations.

For convenience, both Table F-1 and Table F-2 show chemicals designated as HI3 (compared with a TBACT RAL hazard index of 3) and chemicals designated as HI5 (compared with a TBACT level RAL hazard index of 5). See Section 3.1.2.1 for an explanation of how risk should be evaluated for facilities with a mix of HI3 and HI5 chemicals. Appendix A provides example risk evaluation tables for both consideration of target organs, and HI3/HI5 chemicals.

**Table F-1**  
**Chronic Target Organ Systems**

DEQ ID	CASRN/ DEQ ID	Chemical	Noncancer Class		Chronic Target Organ Systems													
			HI3	HI5	Kidney	Liver	Blood	Endo	Musc	Eyes	Skin	Nerv	Cardio	Immune	Resp	Gastro	Develop	Repro
1	75-07-0	Acetaldehyde	HI3												X			
634	67-64-1	Acetone	HI3									X						
3	75-05-8	Acetonitrile	HI3									X			X			
5	107-02-8	Acrolein		HI5											X			
6	79-06-1	Acrylamide	HI3									X						
7	79-10-7	Acrylic acid	HI3												X			
8	107-13-1	Acrylonitrile	HI3												X			
12	107-05-1	Allyl chloride	HI3									X						
13	7429-90-5	Aluminum and compounds		HI5								X						
26	7664-41-7	Ammonia	HI3												X			
30	62-53-3	Aniline		HI5										X				
33	7440-36-0	Antimony and compounds	HI3												X			
37	7440-38-2	Arsenic and compounds	HI3									X						
39	7784-42-1	Arsine	HI3									X						
46	71-43-2	Benzene	HI3				X											
56	100-44-7	Benzyl chloride	HI3												X			
58	7440-41-7	Beryllium and compounds	HI3											X	X			
63	111-44-4	Bis(2-chloroethyl) ether (BCEE)	HI3															
64	542-88-1	Bis(chloromethyl) ether		HI5														
324	74-83-9	Bromomethane (Methyl bromide)	HI3												X			
73	106-94-5	1-Bromopropane (n-propyl bromide)	HI3									X						

**Table F-1**  
**Chronic Target Organ Systems**

DEQ ID	CASRN/ DEQ ID	Chemical	Noncancer Class		Chronic Target Organ Systems													
			HI3	HI5	Kidney	Liver	Blood	Endo	Musc	Eyes	Skin	Nerv	Cardio	Immune	Resp	Gastro	Develop	Repro
75	106-99-0	1,3-Butadiene	HI3															X
333	78-93-3	2-Butanone (Methyl ethyl ketone)	HI3														X	
79	78-92-2	sec-Butyl alcohol	HI3														X	
83	7440-43-9	Cadmium and compounds	HI3		X													
86	105-60-2	Caprolactam	HI3												X			
90	75-15-0	Carbon disulfide	HI3									X						
91	56-23-5	Carbon tetrachloride	HI3			X												
92	463-58-1	Carbonyl sulfide	HI3									X						
97	57-74-9	Chlordane	HI3			X												
101	7782-50-5	Chlorine	HI3												X			
102	10049-04-4	Chlorine dioxide	HI3												X			
104	532-27-4	2-Chloroacetophenone		HI5											X			
108	108-90-7	Chlorobenzene	HI3		X													
117	75-68-3	1-Chloro-1,1-difluoroethane	HI3									X						
246	75-45-6	Chlorodifluoromethane (Freon 22)	HI3		X			X										
230	75-00-3	Chloroethane (Ethyl chloride)	HI3						X								X	
118	67-66-3	Chloroform	HI3		X	X												
325	74-87-3	Chloromethane (Methyl chloride)	HI3									X						
130	76-06-2	Chloropicrin	HI3												X			
131	126-99-8	Chloroprene	HI3											X	X			

**Table F-1**  
**Chronic Target Organ Systems**

DEQ ID	CASRN/ DEQ ID	Chemical	Noncancer Class		Chronic Target Organ Systems													
			HI3	HI5	Kidney	Liver	Blood	Endo	Musc	Eyes	Skin	Nerv	Cardio	Immune	Resp	Gastro	Develop	Repro
136	18540-29-9	Chromium VI, chromate and dichromate particulate	HI3												X			
140	7738-94-5	Chromium VI, chromic acid aerosol mist	HI3												X			
146	7440-48-4	Cobalt and compounds	HI3												X			
149	7440-50-8	Copper and compounds	HI3															
152	1319-77-3	Cresols (mixture), including m-cresol, o-cresol, p-cresol	HI3									X						
161	74-90-8	Cyanide, Hydrogen	HI3					X										
162	110-82-7	Cyclohexane	HI3														X	
186	333-41-5	Diazinon	HI3															
190	96-12-8	1,2-Dibromo-3-chloropropane (DBCP)	HI3															X
112	106-46-7	p-Dichlorobenzene (1,4-Dichlorobenzene)	HI3												X			
116	156-60-5	trans-1,2-dichloroethene	HI3															
328	75-09-2	Dichloromethane (Methylene chloride)	HI3			X												
195	78-87-5	1,2-Dichloropropane (Propylene dichloride)	HI3												X			
196	542-75-6	1,3-Dichloropropene	HI3												X			
197	62-73-7	Dichlorvos (DDVP)		HI5								X						
200	200	Diesel Particulate Matter	HI3												X			
201	111-42-2	Diethanolamine	HI3												X			



**Table F-1**  
**Chronic Target Organ Systems**

DEQ ID	CASRN/ DEQ ID	Chemical	Noncancer Class		Chronic Target Organ Systems													
			HI3	HI5	Kidney	Liver	Blood	Endo	Musc	Eyes	Skin	Nerv	Cardio	Immune	Resp	Gastro	Develop	Repro
271	110-49-6	Ethylene glycol monomethyl ether acetate	HI3															X
236	75-21-8	Ethylene oxide	HI3									X						
239	<a href="#">239</a>	Fluorides	HI3						X									
241	7782-41-4	Fluorine gas	HI3															
250	50-00-0	Formaldehyde	HI3												X			
254	111-30-8	Glutaraldehyde		HI5											X			
286	77-47-4	Hexachlorocyclopentadiene	HI3												X			
287	67-72-1	Hexachloroethane	HI3									X						
297	822-06-0	Hexamethylene-1,6-diisocyanate		HI5											X			
289	110-54-3	Hexane	HI3									X						
290	302-01-2	Hydrazine	HI3			X												
292	7647-01-0	Hydrochloric acid	HI3												X			
240	7664-39-3	Hydrogen fluoride	HI3						X									
293	7783-06-4	Hydrogen sulfide	HI3												X			
300	78-59-1	Isophorone	HI3														X	
302	67-63-0	Isopropyl alcohol	HI3															X
157	98-82-8	Isopropylbenzene (Cumene)	HI3		X			X										
305	7439-92-1	Lead and compounds	HI3									X					X	
311	108-31-6	Maleic anhydride		HI5											X			
312	7439-96-5	Manganese and compounds	HI3									X					X	
316	7439-97-6	Mercury and compounds	HI3									X						



**Table F-1**  
**Chronic Target Organ Systems**

DEQ ID	CASRN/ DEQ ID	Chemical	Noncancer Class		Chronic Target Organ Systems													
			HI3	HI5	Kidney	Liver	Blood	Endo	Musc	Eyes	Skin	Nerv	Cardio	Immune	Resp	Gastro	Develop	Repro
321	67-56-1	Methanol	HI3														X	
329	101-77-9	4,4'-Methylenedianiline (and its dichloride)		HI5						X								
337	108-10-1	Methyl isobutyl ketone (MIBK, Hexone)	HI3														X	
299	624-83-9	Methyl isocyanate	HI3												X			
339	80-62-6	Methyl methacrylate		HI5											X			
346	1634-04-4	Methyl tert-butyl ether	HI3		X					X								
298	101-68-8	Methylene diphenyl diisocyanate (MDI)	HI3												X			
428	91-20-3	Naphthalene	HI3									X			X			
365	365	Nickel compounds, insoluble	HI3											X	X			
368	368	Nickel compounds, soluble	HI3											X	X			
377	7697-37-2	Nitric acid		HI5														
381	98-95-3	Nitrobenzene	HI3												X			
389	79-46-9	2-Nitropropane	HI3			X												
589	8014-95-7	Oleum (fuming sulfuric acid)	HI3															
446	56-38-2	Parathion	HI3															
497	108-95-2	Phenol	HI3			X						X						
503	75-44-5	Phosgene	HI3												X			
506	7803-51-2	Phosphine	HI3		X		X					X			X	X		
507	7664-38-2	Phosphoric acid	HI3												X			
636	12185-10-3	Phosphorus, white	HI3												X			
525	85-44-9	Phthalic anhydride	HI3							X					X			

**Table F-1**  
**Chronic Target Organ Systems**

DEQ ID	CASRN/ DEQ ID	Chemical	Noncancer Class		Chronic Target Organ Systems													
			HI3	HI5	Kidney	Liver	Blood	Endo	Musc	Eyes	Skin	Nerv	Cardio	Immune	Resp	Gastro	Develop	Repro
447	447	Polybrominated diphenyl ethers (PBDEs)	HI3															
645	645	<b>Polychlorinated biphenyls (PCBs) TEQ</b>	HI3			X	X							X	X			
463	32598-13-3	<i>PCB 77 [3,3',4,4'-tetrachlorobiphenyl]</i>	HI3			X	X							X	X			
464	70362-50-4	<i>PCB 81 [3,4,4',5-tetrachlorobiphenyl]</i>	HI3			X	X							X	X			
466	32598-14-4	<i>PCB 105 [2,3,3',4,4'-pentachlorobiphenyl]</i>	HI3			X	X							X	X			
467	74472-37-0	<i>PCB 114 [2,3,4,4',5-pentachlorobiphenyl]</i>	HI3			X	X							X	X			
468	31508-00-6	<i>PCB 118 [2,3',4,4',5-pentachlorobiphenyl]</i>	HI3			X	X							X	X			
469	65510-44-3	<i>PCB 123 [2,3',4,4',5'-pentachlorobiphenyl]</i>	HI3			X	X							X	X			
470	57465-28-8	<i>PCB 126 [3,3',4,4',5-pentachlorobiphenyl]</i>	HI3			X	X							X	X			
474	38380-08-4	<i>PCB 156 [2,3,3',4,4',5-hexachlorobiphenyl]</i>	HI3			X	X							X	X			
475	69782-90-7	<i>PCB 157 [2,3,3',4,4',5'-hexachlorobiphenyl]</i>	HI3			X	X							X	X			
476	52663-72-6	<i>PCB 167 [2,3',4,4',5,5'-hexachlorobiphenyl]</i>	HI3			X	X							X	X			
477	32774-16-6	<i>PCB 169 [3,3',4,4',5,5'-hexachlorobiphenyl]</i>	HI3			X	X							X	X			
481	39635-31-9	<i>PCB 189 [2,3,3',4,4',5,5'-heptachlorobiphenyl]</i>	HI3			X	X							X	X			

**Table F-1**  
**Chronic Target Organ Systems**

DEQ ID	CASRN/ DEQ ID	Chemical	Noncancer Class		Chronic Target Organ Systems													
			HI3	HI5	Kidney	Liver	Blood	Endo	Musc	Eyes	Skin	Nerv	Cardio	Immune	Resp	Gastro	Develop	Repro
646	646	Polychlorinated dibenzo- p-dioxins (PCDDs) & dibenzofurans (PCDFs) TEQ	HI3			X	X							X	X			
527	1746-01-6	2,3,7,8-Tetrachlorodibenzo- p-dioxin (TCDD)	HI3			X	X							X	X			
528	40321-76-4	1,2,3,7,8- Pentachlorodibenzo-p- dioxin (PeCDD)	HI3			X	X							X	X			
529	39227-28-6	1,2,3,4,7,8- Hexachlorodibenzo-p- dioxin (HxCDD)	HI3			X	X							X	X			
530	57653-85-7	1,2,3,6,7,8- Hexachlorodibenzo-p- dioxin (HxCDD)	HI3			X	X							X	X			
531	19408-74-3	1,2,3,7,8,9- Hexachlorodibenzo-p- dioxin (HxCDD)	HI3			X	X							X	X			
532	35822-46-9	1,2,3,4,6,7,8- Heptachlorodibenzo-p- dioxin (HpCDD)	HI3			X	X							X	X			
533	3268-87-9	Octachlorodibenzo-p- dioxin (OCDD)	HI3			X	X							X	X			
539	51207-31-9	2,3,7,8- Tetrachlorodibenzofuran (TcDF)	HI3			X	X							X	X			
540	57117-41-6	1,2,3,7,8- Pentachlorodibenzofuran (PeCDF)	HI3			X	X							X	X			

**Table F-1**  
**Chronic Target Organ Systems**

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**Table F-1**  
**Chronic Target Organ Systems**

DEQ ID	CASRN/ DEQ ID	Chemical	Noncancer Class		Chronic Target Organ Systems													
			HI3	HI5	Kidney	Liver	Blood	Endo	Musc	Eyes	Skin	Nerv	Cardio	Immune	Resp	Gastro	Develop	Repro
563	75-56-9	Propylene oxide	HI3												X			
572	<a href="#">572</a>	Refractory Ceramic Fibers		HI5											X			
577	7783-07-5	Selenide, hydrogen	HI3															
575	7782-49-2	Selenium and compounds	HI3															
579	7631-86-9	Silica, crystalline (respirable)		HI5											X			
582	1310-73-2	Sodium hydroxide	HI3															
585	100-42-5	Styrene	HI3									X						
588	505-60-2	Sulfur Mustard	HI3															
590	7446-11-9	Sulfur trioxide		HI5											X			
591	7664-93-9	Sulfuric acid		HI5											X			
488	127-18-4	Tetrachloroethene (Perchloroethylene)	HI3									X						
245	811-97-2	1,1,1,2-Tetrafluoroethane	HI3															X
599	7550-45-0	Titanium tetrachloride	HI3												X			
600	108-88-3	Toluene	HI3									X						
601	26471-62-5	Toluene diisocyanates (2,4- and 2,6-)	HI3												X			
326	71-55-6	1,1,1-Trichloroethane (Methyl chloroform)	HI3			X												
608	79-01-6	Trichloroethene (TCE, Trichloroethylene)	HI3											X			X	
609	96-18-4	1,2,3-Trichloropropane		HI5											X			
610	121-44-8	Triethylamine	HI3			X				X					X			
613	526-73-8	1,2,3-Trimethylbenzene	HI3									X						

**Table F-1**  
**Chronic Target Organ Systems**

DEQ ID	CASRN/ DEQ ID	Chemical	Noncancer Class		Chronic Target Organ Systems													
			HI3	HI5	Kidney	Liver	Blood	Endo	Musc	Eyes	Skin	Nerv	Cardio	Immune	Resp	Gastro	Develop	Repro
614	95-63-6	1,2,4-Trimethylbenzene	HI3									X						
615	108-67-8	1,3,5-Trimethylbenzene	HI3									X						
620	7440-62-2	Vanadium (fume or dust)	HI3												X			
621	1314-62-1	Vanadium pentoxide	HI3												X			
622	108-05-4	Vinyl acetate	HI3												X			
623	593-60-2	Vinyl bromide		HI5		X												
624	75-01-4	Vinyl chloride	HI3			X												
627	75-35-4	Vinylidene chloride	HI3			X												
628	1330-20-7	Xylene (mixture), including m-xylene, o-xylene, p- xylene	HI3							X		X			X			

**Notes for Table F-1:**

CASRN = Chemical Abstracts Service Registry Number

If no CASRN is available, DEQ identification number is provided.

Endo = Endocrine system

Musc = Musculo-skeletal system

Nerv = Nervous system

Cardio = Cardiovascular system

Immune = Immune system

Resp = Respiratory system

Gastro = Gastrointestinal system

Develop = Developmental effects

**Table F-1  
Chronic Target Organ Systems**

DEQ ID	CASRN/ DEQ ID	Chemical	Noncancer Class		Chronic Target Organ Systems												
			HI3	HI5	Kidney	Liver	Blood	Endo	Musc	Eyes	Skin	Nerv	Cardio	Immune	Resp	Gastro	Develop

Repro = Reproductive effects

**Bold text**= Category name

*Italic text* = Chemical within a category

HI3 = chemical assigned a noncancer Toxics Best Available Control Technology Risk Action Level hazard index of 3.

HI5 = chemical assigned a noncancer Toxics Best Available Control Technology Risk Action Level hazard index of 5.

**Table F-2**  
**Acute Target Organ Systems**

[illegible]



**Table F-2**  
**Acute Target Organ Systems**

[illegible]



**Table F-2**  
**Acute Target Organ Systems**

DEQ ID	CASRN/ DEQ ID	Chemical	Noncancer Class		Acute Target Organ Systems													
			HI3	HI5	Kidney	Liver	Blood	Endo	Musc	Eyes	Skin	Nerv	Cardio	Immune	Resp	Gastro	Develop	Repro
229	100-41-4	Ethyl benzene	HI3									X						
232	106-93-4	Ethylene dibromide (EDB, 1,2-Dibromoethane)	HI3															
233	107-06-2	Ethylene dichloride (EDC, 1,2-Dichloroethane)	HI3															
234	107-21-1	Ethylene glycol	HI3												X			
267	111-76-2	Ethylene glycol monobutyl ether	HI3				X											
268	110-80-5	Ethylene glycol monoethyl ether	HI3														X	
269	111-15-9	Ethylene glycol monoethyl ether acetate	HI3														X	
270	109-86-4	Ethylene glycol monomethyl ether	HI3														X	
271	110-49-6	Ethylene glycol monomethyl ether acetate	HI3															
236	75-21-8	Ethylene oxide	HI3		X													
239	<a href="#">239</a>	Fluorides	HI3												X			
241	7782-41-4	Fluorine gas	HI3							X					X			
250	50-00-0	Formaldehyde	HI3												X			
254	111-30-8	Glutaraldehyde		HI5											X			
286	77-47-4	Hexachlorocyclopentadiene	HI3												X			
287	67-72-1	Hexachloroethane	HI3									X						
297	822-06-0	Hexamethylene-1,6-diisocyanate		HI5											X			
289	110-54-3	Hexane	HI3															
290	302-01-2	Hydrazine	HI3			X												
292	7647-01-0	Hydrochloric acid	HI3												X			
240	7664-39-3	Hydrogen fluoride	HI3												X			
293	7783-06-4	Hydrogen sulfide	HI3												X			

**Table F-2**  
**Acute Target Organ Systems**

DEQ ID	CASRN/ DEQ ID	Chemical	Noncancer Class		Acute Target Organ Systems													
			HI3	HI5	Kidney	Liver	Blood	Endo	Musc	Eyes	Skin	Nerv	Cardio	Immune	Resp	Gastro	Develop	Repro
300	78-59-1	Isophorone	HI3															
302	67-63-0	Isopropyl alcohol	HI3							X					X			
157	98-82-8	Isopropylbenzene (Cumene)	HI3															
305	7439-92-1	Lead and compounds	HI3									X					X	
311	108-31-6	Maleic anhydride		HI5														
312	7439-96-5	Manganese and compounds	HI3									X					X	
316	7439-97-6	Mercury and compounds	HI3									X					X	
321	67-56-1	Methanol	HI3									X						
329	101-77-9	4,4'-Methylenedianiline (and its dichloride)		HI5														
337	108-10-1	Methyl isobutyl ketone (MIBK, Hexone)	HI3															
299	624-83-9	Methyl isocyanate	HI3															
339	80-62-6	Methyl methacrylate		HI5														
346	1634-04-4	Methyl tert-butyl ether	HI3									X						
298	101-68-8	Methylene diphenyl diisocyanate (MDI)	HI3												X			
428	91-20-3	Naphthalene	HI3												X			
365	365	Nickel compounds, insoluble	HI3											X				
368	368	Nickel compounds, soluble	HI3											X				
377	7697-37-2	Nitric acid		HI5											X			
381	98-95-3	Nitrobenzene	HI3															
389	79-46-9	2-Nitropropane	HI3															
589	8014-95-7	Oleum (fuming sulfuric acid)	HI3												X			
446	56-38-2	Parathion	HI3									X						
497	108-95-2	Phenol	HI3							X					X			

**Table F-2**  
**Acute Target Organ Systems**

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**Table F-2**  
**Acute Target Organ Systems**

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**Table F-2**  
**Acute Target Organ Systems**

DEQ ID	CASRN/ DEQ ID	Chemical	Noncancer Class		Acute Target Organ Systems													
			HI3	HI5	Kidney	Liver	Blood	Endo	Musc	Eyes	Skin	Nerv	Cardio	Immune	Resp	Gastro	Develop	Repro
542	70648-26-9	1,2,3,4,7,8-Hexachlorodibenzofuran (HxCDF)	HI3															
543	57117-44-9	1,2,3,6,7,8-Hexachlorodibenzofuran (HxCDF)	HI3															
544	72918-21-9	1,2,3,7,8,9-Hexachlorodibenzofuran (HxCDF)	HI3															
545	60851-34-5	2,3,4,6,7,8-Hexachlorodibenzofuran (HxCDF)	HI3															
546	67562-39-4	1,2,3,4,6,7,8-Heptachlorodibenzofuran (HpCDF)	HI3															
547	55673-89-7	1,2,3,4,7,8,9-Heptachlorodibenzofuran (HpCDF)	HI3															
548	39001-02-0	Octachlorodibenzofuran (OCDF)	HI3															
406	50-32-8	Benzo[a]pyrene	HI3														X	X
559	123-38-6	Propionaldehyde		HI5														
561	115-07-1	Propylene		HI5														
562	6423-43-4	Propylene glycol dinitrate		HI5								X						
273	107-98-2	Propylene glycol monomethyl ether	HI3															
563	75-56-9	Propylene oxide	HI3												X			
572	<a href="#">572</a>	Refractory Ceramic Fibers		HI5														
577	7783-07-5	Selenide, hydrogen	HI3							X					X			
575	7782-49-2	Selenium and compounds	HI3							X	X	X			X			

**Table F-2**  
**Acute Target Organ Systems**

[illegible]



**Table F-2**  
**Acute Target Organ Systems**

DEQ ID	CASRN/ DEQ ID	Chemical	Noncancer Class		Acute Target Organ Systems													
			HI3	HI5	Kidney	Liver	Blood	Endo	Musc	Eyes	Skin	Nerv	Cardio	Immune	Resp	Gastro	Develop	Repro
627	75-35-4	Vinylidene chloride	HI3			X												
628	1330-20-7	Xylene (mixture), including m-xylene, o-xylene, p-xylene	HI3									X			X			

**Notes for Table F-2:**

CASRN = Chemical Abstracts Service Registry Number

If no CASRN is available, DEQ identification number is provided.

Endo = Endocrine system

Musc = Musculo-skeletal system

Nerv = Nervous system

Cardio = Cardiovascular system

Immune = Immune system

Resp = Respiratory system

Gastro = Gastrointestinal system

Develop = Developmental effects

Repro = Reproductive effects

**Bold text** = Category name

*Italic text* = Chemical within a category

HI3 = chemical assigned a noncancer Toxics Best Available Control Technology Risk Action Level hazard index of 3.

HI5 = chemical assigned a noncancer Toxics Best Available Control Technology Risk Action Level hazard index of 5.

# APPENDIX G

## Handling of Non-Detect Values in Risk Assessment

For source tests and ambient monitoring, it is possible that some of the toxic air contaminants will not be detected above the method detection limit. The following is DEQ's preferred approach for handling non-detect values in risk assessments. For calculations involving toxicity equivalency factors, follow DEQ's Source Sampling Manual (Vol. 1, January, 1976, revised November, 2018, <https://www.oregon.gov/deq/FilterDocs/SSMI.pdf>).

### G.1 Source Testing

When considering whether source test data provided by a facility should be accepted for use in a Cleaner Air Oregon risk assessment, DEQ's expectations are the following:

- 1) Source testing will comply with DEQ's Source Sampling Manual; and
- 2) The toxic air contaminants to be tested for, test methods, test conditions, and detection limits will be approved by a DEQ source test coordinator prior to conducting the test; and
- 3) Non-detects will show the actual analytical limit of detection for each source test run; and
- 4) The source test was performed less than two years prior to submittal of the CAO emissions inventory, unless the facility demonstrates that earlier test data remain valid; and
- 5) For cyclic operations or variations in feedstock, tests are representative of variations in loads, feed rates, and seasons, if applicable. An adequate number of tests must be conducted for all cyclic or seasonal operations.

DEQ's Source Sampling Manual requires that the detection limit be used for non-detect values in averaging data, except that substitutions at less than the detection limit may be used in CAO risk assessments if approved by DEQ. Provided the above conditions are met, DEQ will accept the following:

**If a toxic air contaminant is not detected in any source test runs or samples**, you can consider the toxic air contaminant not present, and treat its concentration as zero in that portion of the risk assessment.

**If a toxic air contaminant is detected in less than 10% of the test runs or samples**, assign a concentration of **zero** to those test runs or samples that were non-detect. Average the non-detect values (zeros) with the detected values, and report the final average value for use in the risk assessment.

**If a toxic air contaminant is detected in 10% or more of the test runs or samples**, assign a concentration of **one-half the detection limit** to those test runs or samples that were non-detect. Average the detected values with  $\frac{1}{2}$  detection limit values for the non-detect samples, and report the final average value for use in the risk assessment.

## **G.2 Ambient Monitoring**

### **Non-detects in all test samples**

If all samples consistently show levels below detection limits, consider the toxic air contaminant not present, and treat its concentration as **zero** in that portion of the risk assessment.

### **Non-detects in some test samples**

If toxic air contaminants are detected in some samples but not others, and there are sufficient samples, use an appropriate statistical method such as Kaplan Meier to calculate an exposure concentration. If there are insufficient samples for a meaningful statistical evaluation, handle the non-detect samples consistent with the procedures for source tests listed above.

For both source testing and ambient monitoring, DEQ may make other site-specific decisions regarding non-detect values if, for example, detection limits are found to be unreasonably high after DEQ approved a monitoring plan with reasonable detection limits.